

## Manuscript Details

<b>Manuscript number</b>	PEDOBI_2017_124_R1
<b>Title</b>	ASSESSMENT OF SOIL MICROBIAL FUNCTIONAL DIVERSITY: LAND USE AND SOIL PROPERTIES AFFECT CLPP-MICRORESP AND ENZYMES RESPONSES
<b>Article type</b>	Research Paper

### Abstract

The assessment of microbial functional diversity is an important indicator of soil quality. Different methodological approaches are currently used; among them are enzyme activities (EA) and CLPP (community level physiological profile) techniques (e.g. MicroResp<sup>TM</sup>, MR). The aims of the study were: i) to assess the efficacy of both methods in capturing differences among various land use categories when different levels of selected explanatory variables such as land use category, total organic carbon (TOC) and pH are considered, and ii) to explore, through a quantile regression approach, the possible relationships between each of the two methods with land use category, TOC and pH. The Shannon diversity index ( $H'$ ), calculated from EA and MR data, was chosen as a synthetic index deriving from the same mathematical model. The quantile regression model (QRM), the Kruskal-Wallis and Spearman rank correlation tests were performed. Enzyme activities and MicroResp were reliable ecological indicators to assess soil microbial functional diversity. No correlation was found between the diversity indexes,  $H'$ EA and  $H'$ MR, it was therefore supposed that the two methods may target complementary components of microbial functional diversity. Both methods were effective in capturing differences among various land use categories, in particular  $H'$ MR in soils with low TOC content (<1.5%). Moreover, the QRM approach allowed a more detailed analysis along the distribution of the diversity indexes ( $H'$ EA and  $H'$ MR) indicating that  $H'$ EA was more dependent on the selected variables.

<b>Keywords</b>	Microbial processes; Shannon index; Soil properties.
<b>Manuscript category</b>	Soil microbial ecology
<b>Corresponding Author</b>	Sara Marinari
<b>Order of Authors</b>	Maria Cristina Moscatelli, Luca Secondi, Ruxandra Papp, Rosita Marabottini, Stazi Silvia Rita, Elena Mania, Sara Marinari
<b>Suggested reviewers</b>	José Luis Moreno-Ortego, Teresa Hernandez

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Figures.docx [Figure]

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## Research Data Related to this Submission

There are no linked research data sets for this submission. The following reason is given:  
We are able to provide some data upon request. However, since the work has been written analysing several different case studies, when these were not published, authorization needs to be asked

Viterbo October 30th, 2017

To the Editor of  
Pedobiologia

Manuscript n. PEDOBI\_2017\_124

Dear Editor,

We wish to submit the revised article entitled "**ASSESSMENT OF SOIL MICROBIAL FUNCTIONAL DIVERSITY: LAND USE AND SOIL PROPERTIES AFFECT CLPP-MICRORESP AND ENZYMES RESPONSES**" modified following both yours and reviewers' comments.

The paper has been deeply revised in most of its sections in relation to the new results obtained after standardization of data asked by reviewer n. 1 and further amendments asked by reviewer #2.

Therefore also the discussion and conclusions were focused on the new aspects emerged.

For this reason we also agreed to change the title as you suggested.

We deeply thank you and the two anonymous reviewers for all the precious comments very useful to improve the whole paper.

We declare that we do not have conflicts of interest to disclose.

Please address all correspondence concerning this manuscript to me.

Thank you for your consideration of this manuscript.

Sincerely

Prof. Sara Marinari  
DIBAF - University of Tuscia  
Via S. Camillo de Lellis

**Viterbo** 01100 - ITALY

[marinari@unitus.it](mailto:marinari@unitus.it)

## Comments from the editor and reviewers:

### Editor

The paper has been deeply revised in most of its sections in relation to the new results obtained after standardization of data as asked by reviewer n. 1 and further amendments asked by reviewer #2. Therefore also the discussion and conclusions were focused on the new aspects emerged. For this reason we also agreed to change the title as you suggested into:

**ASSESSMENT OF SOIL MICROBIAL FUNCTIONAL DIVERSITY: LAND USE AND SOIL PROPERTIES AFFECT CLPP-MICRORESP AND ENZYMES RESPONSES**

### -Reviewer 1

#### General comments

The authors perform a meta-analysis of functional diversity measurements (Shannon index) in a range of soils with different land uses and provide a useful comparison of two methodological approaches, which they suggest target different stages in the soil organic matter decomposition process. The values obtained for the index are greater than what I would expect and the calculation should be checked or explained. They apply Quantile Regression Modelling of the data, which they argue is useful for skewed distributions but more discussion is really needed on the meaning of the results. Some figures, or parts thereof, are duplicative and should be omitted. The authors' final conceptual model, while making a valid distinction between the two functional measurements, is rather over-simplified and some of the concepts are a bit strange. In places the English could be improved; some examples are given below but there are others which need some editing.

**We thank reviewer 1 for the general comment. The manuscript has been deeply revised in many of its sections following his/her suggestions and those of reviewer #2.**

#### Specific comments

L16 This first sentence is not actually a sentence; start with something like "Here we consider..."

**Thanks! Actually there was a typing error and "as" was supposed to be "is".**

L18 Replace "i.e." with "e.g."

**Done throughout the manuscript**

L22 "and pH".

**The sentence was changed**

L52 I do not think this is entirely accurate. Catabolic processes continue all the time, generating maintenance energy – they become more apparent when energy for anabolic processes is lacking. Of course, certain types of stress may increase the demand for maintenance.

**The sentence has been removed**

L66 “pedogenic” is better than “pedogenetic”.

**Done**

L92 Replace “cases” with “case”.

**Done**

L93 Replace “On” with “For”.

**Done**

L103 The period of reconditioning should be stated.

**Done**

L121 Macaulay Scientific Consulting Ltd has now been replaced by James Hutton Ltd.

**Done**

L129 It should be explained what the various symbols in the equation refer to and what were the values of the constants (or at least give a reference).

**The conversion of absorbance to % CO<sub>2</sub> is a non linear relationship and the best fitted curve (regression analysis) is used to obtain the formula and parameters. A calibration procedure was performed taking into account the spectrophotometer used, different types of soils and incubation conditions. In our experimental conditions the constants of the following equation  $A+B/(1+ D \times A_i)$  were: A:-1,62, B:-4,85 and D: -8,1. (Campbell et al, 2003). See text.**

L135, 138 I have some difficulty in referring to the Shannon diversity index in terms of entropy. Admittedly Shannon’s work on information theory had its roots in thermodynamics and there are parallels, but when applied to diversity within ecology it becomes a rather different concept. Increase in entropy is seen as negative (e.g. heat death of the universe), while increase in diversity is usually seen as something positive. In the context of the paper, I am not sure what “entropy of a system” and “microbial functions entropy” really mean (yes, I looked at Marinari et al. 2013 but it was not helpful and the reference Minasny et al. 2008 is really talking about the entropy in different mineralogies, with no biological component).

**We agree that the increase of “entropy” of soil microbial functions means something positive, in any case, the term entropy can be also associated to the biodiversity in an ecological context as reported by Spellerberg and Fedor (2003). For this reason we added this reference in the text.**

**Moreover, the reference to Marinari et al. 2013 is related to the fact that microbial functions diversity is linked to the diversity of hydrolysable substrates. However, in that paper the concept of entropy was also related to pedogenesis that leads to a highest energy level when horizons differentiate and the spatial arrangement of soil particles (soil structure) becomes more defined.**

L202 There are no lower outliers for SEI shown in Fig. 1 for any soil group. Also in these ranges, it is more conventional to go from low to high, e.g. 42 to 11821 (the latter should be written as 11800 as 11821 indicates unjustifiable accuracy).

**We modified both the sentence concerning lower outliers and the order of values.**

L206 Figure 2 and Table 2. I am having some difficulty in understanding the range of H' values. The maximum values, and even the mean values, exceed what I would calculate as the maximum possible. Assuming you have used the formula as given in L135, then the maximum value (total evenness) would be 2.708 for 15 'species' (carbon substrates) and 2.079 for 8 'species' (enzyme activities). The data in Table 2 is giving maximum values of 6.720 and 5.490, respectively. Or has the index been calculated in some other way?

**We thank the reviewer for this observation which enabled us to identify some inaccuracies in index calculation. In the updated version of the paper we changed the values and reported descriptive synthesis values in Table 2.**

LL214-216 I am not convinced that Figure 3 is required in addition to Figure 2. It is essentially the same data presented in two ways. Given that the same data is also summarised in Table 2, it does seem to be overkill. However, the contention that H'MR has greater variability is not sustained. To compare properly the two datasets should be standardized – normally by dividing by the mean. The interquartile distance then comes out as 0.24 and 0.25 for the H'EA and H'MR, respectively, – hardly a great difference.

**We removed Figure 3 from the paper. In the revised version of the paper we used standardized measures for both indexes in order to deal with the issue of different ranges. This was specifically indicated in the M&M and Results section.**

LL219-225 It is not necessary to repeat all the values given in Table 3 – omit the  $r_s$  values.

**Done. Only p-values were left in the text.**

L223 At this point there is a switch to sometimes using soil type in place of land use category; it is better to stick with the latter (See also Table 4).

**Done throughout the whole manuscript**

L233ff It is not clear (not being that familiar with quantile regression) what the significance of the constant is, particularly since it seems to be highly significant in all cases. I presume it is just the intercept on the regression but is it the same regardless of whether the regression is against TOC, pH or Soil type (Land Use)?

**The interpretation of intercept in QRM is quite similar to its interpretation in standard linear regression models (OLS) with the utmost importance to keep in mind a different intercept depending on the specific quantile that is being analysed.**

L258ff It could be argued that because only 8 enzymes were assayed in comparison to 15 carbon substrates tested, that the enzyme approach in gauging diversity was necessarily more limited. Do you have a counter to this suggestion?

**We agree that Shannon index is calculated using different numbers of enzymes and substrates (8 enzymes belonging to the 4 nutrients –C,N,P,S - biogeochemical cycles and 15 substrates representing four ecologically relevant categories of biomolecules – proteins, carbohydrates, organic acids and phenols). However, the aim of the study was to assess the efficacy of the two techniques, as they are generally used in the literature, to calculate a synthetic index aimed to capture differences among the different land uses when different levels of pH and TOC are considered. In particular we would like to stress that this study is a meta-analysis that has been conducted using data provided by previous researches performed. However, the comment is proper and we agree that to promote the use of both techniques in the same study, and to improve the interpretation of the obtained results, it should be suggested to select the same number of enzymes and substrates.**

L280ff At the end of the day both enzyme activity (as assayed) and CLPP are both degradative, just that the former is one step back in the chain of events. One might have expected a greater degree of correlation in H' values. However, H' is only one way of expressing/summarising the data. What would have been the result had you looked at total activity (Figure 1) and made a comparison? Was this done or is it the subject of a separate study?

**We thank the reviewer for this comment. The correlation between the functional capacity (total activity) measured according to the two methods (SEI and SIR reported in figure 1) showed a significant coefficient ( $p < 0.05$ ). We would like to stress that even if, the functional capacity measured by means of the two methods was correlated, the functional diversity was not. This means that, although the capability to perform functions by enzyme and respiration were positively related, their variability (diversity of functions) was not.**

LL302-303. Is this statement true? Admittedly immobilized enzymes can have little impact on solid substrates such as cellulose and hemicellulose because of spatial separation, but once polymeric fragments are solubilised such enzymes can then come into contact with them. Their monomeric products then become available to the microbial community. Your tests with MUF substrates demonstrates that immobilized enzymes are still active against low molecular weight intermediates (if indeed this is the case).

**Yes, the statement is true and we do not understand what the reviewer referred to. The sentence at lines 302-303 referred to the fact that, being enzymes also in the soil in the immobilized forms, these may not be directly expression of microbial activity, thus of microbial functional diversity. Immobilized enzymes represent a background biological activity giving resilience to soils under unfavourable conditions for microbial life. However, due to methodological limitations, we cannot subtract the immobilized enzymes contribution to the total activity measured in the laboratory with the current available methods.**

LL311-315 There is not much discussion on the QRM results. Quite a lot of space is devoted to the methodology and results of the QRM so I was hoping for a bit more explanation as to what the ecological implications of the findings were.

**We reduced the theoretical explanation of QRM consistently to what is required by Reviewer 2 as well. Moreover we added explanation concerning ecological implications of the results.**

L315 Spelling of Zalnina?

**Corrected**

Figure 3. The H'MR result does not need to be dashed – not done for other figures.

**In accordance with your comments concerning Figure 3 (LL214-216) we removed this figure from the paper.**

Figure 4. The two parts (figure and table) duplicate each other. The figure part should be omitted since all the information is in the table. It would probably aid clarity if numbers are given to three significant figures only (greater accuracy is unwarranted).

**We deleted Figure 4; in the revised version of the paper only the table is reported, now table 4. Moreover, we considered three significant figures.**

Table 1 L3 the abbreviation is “conv” not “con”.

**Done**

Table 3 This table seems to be overly complicated. Why not 1X4 in place of 4X4, i.e. the four values in one row?

**The table has been simplified as suggested**

Table 4 TOC and Soil type are given as discrete variables whereas pH is given as a single (continuous?) variable, not as the ranges given in Figure 4 – this seems to be rather inconsistent. In actuality, for the purpose of these regressions, would it not be better to treat TOC as the continuous variable it is, rather than boxing it into these three categories?

**We re-estimated QRM by considering pH in classes (as Table 5). The distinction of variables into classes enabled us to detect significant changes of the relationship with the dependent variables and within/across quantiles.**

## Reviewer 2

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### General Comments

Authors present an interesting study about the functional diversity of soil microbial activity in a wide range of soils submitted to different utilization. The variation of Shannon's Index of diversity for microbial functional diversity was studied by measuring the profiles both of different enzyme activities and micro-respiration of diverse C substrates. Also they determine the effect of pH and TOC of soil in this microbial functional diversity. Important conclusions are derived from the obtained results in this study. In general the manuscript is written in a good English language, although some specific parts of the text should be re-written in a clearly and simply way.

### Specific comments

#### Abstract

L16. In this sentence, a verb is missing. Please re-write in a grammatically correct form.

**The sentence was corrected, actually there was a typing error and "as" was supposed to be "is".**

L22. Insert the conjunction "and" between the words "category" and "pH".

**The sentence has been changed**

L74. Correct publication year: 2002.

#### Done

#### Material and Methods

This section should be structured in several sub-sections, e.g.: Experimental design, soil sampling, soil analysis methodologies, statistical methods...

#### Done

L92. "Cases studies" should be write "case studies".

#### Done

What do you mean? Are these case studies treatments or conditions? Explain in a clearly way.

**With the term "case study" we mean a particular investigation that included different treatments. For example: one of the 4 case studies included in the "forest soils category F" was related to different management practices in two adjacent soils, one under native forest and the other one under a recently coppiced forest. This was clarified in the text.**

L95. Delete colon (:)

#### Deleted

L96. Include conjunction "and" between the words "afforestation" and "chronosequences".

#### Done

L98. Firstly, substitute conjunction "and" by comma and the include conjunction "and" between the words "tillage" and "natural".



Done

L100. Insert conjunction "and" between the words "paddies" and "highly".

Done

L101. Explain better the case studies of EC category because only five case studies are distinguished.

**There was a typing error. The case studies belonging the EC category are 6: three of them subjected to natural pedoclimatic conditions and the other three under heavy anthropic impact. It has been better specified in the text.**

L103. You should specify conservation or store method of soil samples.

**Conservation and treatment of soils from sampling to analyses has been detailed in the text.**

L111. Why did you only analyze the acid-phosphatase activity? Alkaline phosphatase activity is a more important activity in alkaline soils.

**We agree with this observation. However, in all case studies the same experimental set-up for enzymatic assays has been used and this allowed to perform this meta-analysis. Eight different enzymes on several soil samples were measured at the same time in the same microplate using a common buffer as that suggested by Marx et al. (2001) (NaAc pH 5,5). In this way the biondicator used responds to the requisites of providing fast results while processing a great amount of soils in a short time.**

L129. If you include this equation you should explain what is every variable or parameter of it.

**A calibration procedure was performed taking into account the spectrophotometer used, different types of soils and incubation conditions. In our experimental conditions the constants of the following equation  $A+B/(1+ D \times Ai)$  were: A:-1,62, B:-4,85 and D: -8,1. (Campbell et al, 2003). See text**

L131. Please explain SEI and SIR and how they are calculate or measured. Dumontet et al., 2001 is not a pertinent reference for specifically citation about Synthetic Enzymatic Index, because in this study this index is not introduced.

**Explanation of calculation of SEI and SIR has been given in the M&M section. Dumontet et al., (2001) provided the suggestion of combining some enzyme activities leading to the same final product (e.g. pNP or MUF) as a synthetic index. However, we removed this citation since it is considered not pertinent.**

L152. Specify these two distributions because Figure 1 represents SEI and SIR values and Figure 2 truly shows the distributions of H'EA and H'MR.

**Modified. In the revised version of the paper we specified that this sentence refers to the two distributions illustrated in Figure 2.**

L154. Re-write this sentence, clearly specifying the two explanatory variables.

**The sentence has been clarified presenting the selected covariates in this study**

L168-186. The explanation of the quantile regression model used in this study should be re-written in a clearly and simply way, without mathematical equations. Authors are not expected to write a statistical treatise in this part of the manuscript. Authors should explain what does it consist in and why do you use this particular regression?

**This section has been re-written without mathematical equations and emphasizing the advantages of QRMs in soil analysis.**

L192-197. In this paragraph, you should write statistical significance of the relationship. The correlation coefficient establishes the relationship level between two variables.

**We added the level of significance.**

*Results*

L202. Delete "and lower". Only upper outliers are shown in figure 1a

**Modified**

L201-204. Re-write this sentence in a clearly and simply way.

**Modified, taking into consideration also suggestions by Reviewer 1.**

L210. Delete "the null hypothesis of normal data is rejected for both distributions". This part of the sentence is obvious and reiterative because it was previously stated that these two distributions were significantly different from normality.

**Done**

L217-218. Delete "lower" and "upper" because the outliers are shown under and above the whiskers.

**We deleted box-plots taking also into consideration suggestions by Reviewer 1**

L225. Authors should include p-values for these two cases.

**Modified**

L226. Delete "significantly" and after "distinguished" add "in different ways".

**Modified**

L227. Substitute "However" by "Thus".

**Modified**

L239. Write "quantile" in singular.

**Modified**

L243. After "relationship", include "with the land use category A".

**Comments have been modified according to the new estimated models.**

L245. "this land use category A". Re-write such as: "the land use category A"

**Comments have been modified according to the new estimated models.**

L258-267. In this paragraph, authors should include an explanation how pH can affect microbial function diversity represented by H'MR.

**Done**

L324. Delete the preposition "at".

**Done**

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**LEGENDA:**

**Yellow marked text: modified text**

**Blu marked text: deleted text**

**Comments mark where new sentences have been added**

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**EFFECTIVENESS OF ENZYME ACTIVITIES AND CLPP-MICRORESP AS**

**INDICATORS OF MICROBIAL FUNCTIONAL DIVERSITY IN A WIDE RANGE OF**

**SOILS**

**Moscatelli M.C.<sup>1</sup>, Secondi L.<sup>1</sup>, Marabottini R.<sup>1</sup>, Papp R.<sup>1</sup>, Stazi S.R.<sup>1</sup>, Mania E.<sup>2</sup>, Marinari S.\*<sup>1</sup>**

<sup>1</sup>*Dept. for Innovation in Biological, Agrofood and Forest systems (DIBAF), University of Tuscia, Viterbo, Italy*

<sup>2</sup>*Department of Agricultural and Forest Sciences (DISAFA), University of Torino, Torino, Italy*

\*corresponding author [marinari@unitus.it](mailto:marinari@unitus.it), tel. 0039-(0)761-357288

**Keywords: microbial processes, land-use, Shannon index, quantile regression model**

## 22 **Abstract**

23 The assessment of microbial functional diversity, as an important indicator of soil quality. Different  
24 methodological approaches are currently used; among them are enzyme activities (EA) and CLPP  
25 (community level physiological profile) techniques (i.e. MicroResp<sup>TM</sup>, MR). The aims of the study  
26 were: *i*) to assess the efficacy of both methods in capturing differences among various types of soils  
27 when different levels of selected explanatory variables such as Total Organic Carbon (TOC), land use  
28 and pH are considered, and *ii*) to explore, through a quantile regression approach, the possible  
29 relationships between each of the two methods with TOC, land use category, pH. TOC and pH were  
30 chosen as explanatory variables influencing microbial functional diversity. The Shannon diversity  
31 index ( $H'$ ), calculated from EA and MR data, was chosen as a synthetic index deriving from the same  
32 mathematical model. The quantile regression model (QRM), the Kruskal-Wallis and Spearman rank  
33 correlation tests were performed. The QRM and Kruskal-Wallis tests evidenced that MicroResp  
34 technique generally provided a higher discrimination capacity within different land use categories,  
35 TOC and pH ranges (TOC <0.15-8.41%; pH <4.02-9.01>). Soil pH was found to be a key property,  
36 rather than TOC content, in differentiating microbial processes.  $H'$ EA and  $H'$ MR were not correlated  
37 but, when analysed separately, only agricultural soils showed a weak correlation ( $P < 0.1$ ) probably  
38 due to the fact that these soils features fall within the intermediate range of pH and TOC where both  
39 methods were found to be significantly sensitive. These results suggest that the two methodologies do  
40 not target the same microbial processes. We hypothesize that the two methodologies refer to  
41 sequential steps of microbial activity. In fact, pointing to complementary components of microbial  
42 functional diversity EA and MR provide a different ecological significance which may inform on the  
43 extent of dissipating energy pathways in the soil system.

## 44 **Introduction**

45 The links between ecosystem functioning and levels of soil biodiversity have been the focus of the  
46 recent scientific literature (Delgado-Baquerizo et al. 2016; Griffiths et al. 2016; Nannipieri et al.  
47 2003). The first authors provided evidence that loss in microbial diversity will likely reduce multiple  
48 ecosystems functions thus negatively impacting the provision of ecosystem services. Adhikari &  
49 Hartemink (2016) claimed for new insights into soil microbial diversity and their role in soil  
50 functional variability. Since up to 80/90% of soil functions, from humification to mineralization, is  
51 microbially-mediated, the diversification of soil microorganisms in terms of structure and/or activity is  
52 essential to maintain functioning of terrestrial ecosystems (Pereira et al. 2013).

53 Microbial functional diversity is defined as “the sum of the ecological processes, and/or capacity to  
54 use different substrates developed by the organisms of a community” (Insam et al. 1989). Emmerling  
55 et al. (2002) and Wellington et al. (2003) report that if microbial genetic diversity assesses a latent  
56 diversity, which may not be expressed, functional diversity is related to the actual activities resulting  
57 from that potential so that "functional rather than taxonomic diversity may provide greater insight to  
58 microbial roles in ecosystems" (Zak et al. 1994). **In fact, under stress or unfavourable conditions,**  
59 **microorganisms may switch from anabolic pathways to catabolic pathways (Anderson and Domsch**  
60 **2010). In this case the soil turns into a dissipating energy system with enhanced energy demand.**

61 Over the last 10 years, the scientific literature provided a great number of papers aimed to assess  
62 microbial functional diversity as an important ecological indicator to monitor and assess soil quality  
63 changes in different pedoclimatic conditions, land uses and human pressure levels (i.e. management  
64 practices)(Bardgett and van der Putten, 2014; Griffiths et al., 2016).

65 To measure the activity and diversity of the microbial community a number of methods can be  
66 applied, to cite few of the most common approaches: (i) catabolic activity investigated by Biolog<sup>TM</sup>-  
67 plates (Garland and Mills, 1991; Rutgers et al. 2016), (ii) respiration of different substrates as  
68 investigated by the MicroResp<sup>TM</sup> method (Campbell et al. 2003; Chapman et al. 2007; Creamer et al.  
69 2016) and (iii) enzyme activities (Nannipieri et al. 2012; Hendriksen et al. 2016).

70 Although all methodological approaches are reliable and sensitive, few studies aimed to understand  
71 their effectiveness to discriminate microbial functional diversity in relation to soil organic C and pH  
72 as the main properties being affected by land use and management practices, anthropic impact and  
73 other **pedogenetic** factors. To achieve this goal, a large number of case studies covering different  
74 types of soils is necessary. In this study, about 200 measurements of microbial functional diversity  
75 obtained over a broad spectrum of key soil properties and across different land uses and management,  
76 were selected. Furthermore, microbial functional diversity obtained through enzyme activities (EA)  
77 and CLPP-MicroResp (MR), was synthetically represented by the Shannon index ( $H'$ ) that  
78 transforms the obtained results to a comparable range of values deriving from the same mathematical  
79 model. The Shannon index is a comprehensive indicator of microbial species, individual numbers and  
80 evenness, or distribution of the enzyme activities and is influenced by richness of community species  
81 (Bending et al. 2004; Li et al., 2007).

82 The aim of the present study was therefore to: i) assess the efficacy of both methods in capturing  
83 differences among the different land use categories when different levels of pH and TOC are  
84 considered, ii) explore, through a quantile regression approach, the possible relationships between  
85 each of the two methods and selected explanatory variables (TOC, land use category, pH).

86 **Furthermore, the results of these analyses could help to assign an ecological significance to both**  
87 **methods in various environmental contexts and research issues.**

88

## 89 **Materials and methods**

90 The results presented in this paper have been obtained performing additional statistical analyses on  
91 data collected in the Laboratory of Chemistry and Biochemistry, University of Tuscia, Viterbo, Italy  
92 during the last 6 years (2010-2016). Microbial functional diversity was measured, by means of  
93 enzyme activities and CLPP-MicroResp<sup>TM</sup> technique, in a wide range of soils analysed within  
94 different research projects. Most of the sampling sites are located within the Mediterranean climatic  
95 area. Other climatic areas are the monsoon one for the Bangladesh case study, the temperate one for

96 Switzerland, oceanic for United Kingdom and boreal for Sweden. All soils represent a broad  
97 spectrum of key soil properties across different land use categories, wide range of soil pH and soil  
98 organic carbon content (TOC) (Table 1).

99 The soils were grouped into three main categories, including 15 cases studies, with the aim to  
100 separate diverse land uses and/or specific conditions. On this purpose, three groups were identified: F  
101 (forest soils, 4 case studies), A (agricultural soils, 5 case studies) and EC (extreme conditions, 6 case  
102 studies). The case studies performed on forest soils (F) included different: management practices,  
103 lithological substrates, afforestation, chronosequences. The soils under agricultural land use (A) were  
104 characterized by different managements and/or agricultural practices such as: organic, biodynamic  
105 and conventional cropping systems, tillage/no tillage, natural green cover/no cover. The third  
106 category (EC) included soils with peculiar characteristics due to pedoclimatic conditions (saline  
107 environments, natural arsenic contamination in rice paddies, highly calcareous soils) or heavy  
108 anthropic impact (a multi-element contaminated dump, arsenic contaminated mine)(Table 1).

109  
110 All soils were sampled at 0-20 cm depth during the dry season (spring/summer), air dried, sieved at 2  
111 mesh and re-conditioned at 60% of their water holding capacity prior to biochemical analyses.

112 The total organic carbon (TOC) was determined by combustion by Shimadzu TOC VCSH analyzer  
113 while soil pH was measured on sieved soil suspended in a solution of deionised water in 1:2.5 ratio  
114 (w/v). The pH was measured in the supernatant with a pH meter (pH 211, Hanna Instruments).

115 A total of 196 values of microbial functional diversity, assessed by means of enzyme activities and  
116 CLPP-MicroResp, were used for this study (Table 1). Enzymes were measured following Marx et al.  
117 (2001) using fluorogenic methylumbelliferyl (MUF)-substrates. Soils were analysed for  
118 cellobiohydrolase,  $\beta$ -1,4-glucosidase,  $\alpha$ -1,4-glucosidase,  $\beta$ -N-acetyl-glucosaminidase,  $\beta$ -1,4-  
119 xylosidase, acid-phosphatase, arylsulphatases and butyrate esterase which is considered a proxy of  
120 endocellular activity (Wittman et al. 2004). The relative fluorogenic substrates, prepared with acetate  
121 buffer 0.5 M pH 5.5, were: 4-MUF- $\beta$ -D-cellobioside, 4-MUF- $\beta$ -D-glucoside, 4-MUF-N-acetyl- $\beta$ -

Commented [1]: Headings included



122 glucosaminide, 4-MUF- $\alpha$ -D-glucoside, 4-MUF-phosphate, 4-MUF-7- $\beta$ -D-xyloside, 4-MUF-sulphate  
123 and 4-MUF-butyrate. Fluorescence (excitation 360 nm, emission 450 nm) was measured with an  
124 automatic fluorimetric plate-reader (Fluoroskan Ascent) and readings were performed after 0, 30, 60,  
125 120 and 180 minutes of incubation at 30° C. The results were expressed as nmoles of product (MUF)  
126 of each enzymatic reaction released per g of soil per unit of time in relation to a standard curve  
127 prepared with increasing MUF concentrations and incubated at the same experimental conditions.

128 The community level physiological profile (CLPP) was determined using the MicroResp™ soil  
129 respiration system (MicroResp™, Macaulay Scientific Consulting Ltd, Aberdeen, UK) according to  
130 Campbell et al. (2003).

131 The 15 substrates selected in this study were:  $\alpha$ -D-glucose, D-Galactose, D-fructose, L-arabinose, L-  
132 leucine, L-arginine, Glycine, L-aspartic acid,  $\gamma$ -amino-butyric and glutamic acid, three carboxylic  
133 acids: citric acid, oxalic acid and L-ascorbic acid, and two phenolic acids: vanillic and syringic acid.

134 The emission of CO<sub>2</sub> by the microbial biomass was estimated using a colorimetric method  
135 (microplate spectrophotometer) before and after 6 h of incubation at 28 °C. The absorbance was read  
136 at 595nm. At the end the absorbance was normalised for any difference recorded at time zero and

137 then converted to % CO<sub>2</sub> using the calibration curve  $y = A+B/(1 + D \times Ai)$ . The CO<sub>2</sub>% was converted  
138 to  $\mu\text{g C-CO}_2 \text{ g}^{-1} \text{ h}^{-1}$  production rate using gas constant, T° C, headspace volume, soil dry weight  
139 (d.w.) and incubation time. The SEI (Synthetic Enzymatic Index, Dumontet et al, 2001) and SIR  
140 (Substrate Induced Respiration) for all soils within the three categories (F, A and EC) have been  
141 calculated as a synthetic measure of microbial functional capacity.

142 Microbial functional diversity was assessed calculating the Shannon-Weaver diversity index  
143 (Kennedy and Smith, 1995) corresponding to the entropy concept defined by:  $H' = - \sum p_i * \ln p_i$

144 (Shannon and Weaver, 1949), where  $p_i$  is the ratio of the activity of a particular enzyme to the sum of  
145 all enzymatic activities (H'EA) or the respiration rate of each single C-substrate for MicroResp™  
146 (H'MR). Shannon diversity index is related to the entropy of a system and when applied as a measure  
147 of microbial functions entropy, may express the heterogeneity of soil organic substrates availability

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148 and microbial processes (Marinari et al., 2013). Since the eight enzymes and the 15 substrates here  
149 tested did show activity in all the analysed samples, then, in this work, the diversity recorded reflects  
150 only the “evenness” or distribution of the enzyme activities or ability to use the different substrates  
151 (Bending et al., 2004; Rodríguez-Loinaz et al., 2008).

152  
153 The analysis of all collected data was carried out into **three main** steps. Firstly, descriptive analyses  
154 provided us with a clear picture of the distribution of the two indexes (H'EA and H'MR) as well as  
155 information about the shape of the two distributions. Moreover, rank correlation measures and test  
156 performed by using the Spearman correlation enabled us to evaluate if, and to what extent, the two  
157 methodological approaches (EA, enzyme activities and MR, MicroResp™) used to evaluate soil  
158 microbial functional diversity are related.

159 The Kruskal-Wallis non-parametric test was used to test if and to what extent the two indexes  
160 distinguished the various land use categories in relation to TOC and pH ranges.

161 By considering the asymmetry of the two distributions (as shown in **Figure 1 and** Figure 2) as well as  
162 the results of the Shapiro-Wilk normality test, we analyzed the existence of association between each  
163 of the two measures and **two explanatory variables, by estimating quantile regression models.**

164 **Indeed, quantile regression offers the possibility to highlight how the effect of the selected covariates**  
165 **changes throughout the entire distribution of the dependent variable.** To estimate the relationships  
166 (**association**) between the dependent variables and the set of selected covariates the classical OLS  
167 (Ordinary Least Squares) regressions can be applied. However, data obtained from experimental  
168 collection tend to be skewed **so** these models **do not** describe the “correct” relationships. **Moreover,**  
169 **from the soil analysis perspective, it is interesting to understand what happens throughout the entire**  
170 **distribution of the two measures and at their extremes.**

171 **Quantile Regression Models (QRMs) are of special interest to studies characterized by skewed**  
172 **distributions. Indeed, these models allow for investigation of the potential different effect of a**  
173 **covariate on various quantiles in the conditional distribution, they are more robust to the presence of**

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174 outliers and can be consistent under weaker stochastic assumption than with least-squares estimation  
 175 (Cameron and Trivedi, 2005). The application of these types of models to our analysis can help to  
 176 understand if and to what extent the differences observed between the two measures can be attributed  
 177 to the different effect played by the explanatory variables at the various quantiles. The QRM  
 178 specifies the conditional quantile of the dependent variable  $y$  as a linear function of covariates  
 179 (Koenker 2005):

$$Q_{\theta}(y_i | \mathbf{x}_i) = \mathbf{x}_i' \boldsymbol{\beta}_{\theta} + \varepsilon_{i\theta} \quad (1)$$

181 where  $y_i$  ( $i=1, \dots, n$ ) is the dependent variable represented, in turn, by  $\mathbf{x}_i$  is a sequence of  $k$ -vector of  
 182 regressors,  $\boldsymbol{\beta}_{\theta}$  is an unknown vector of regression parameters associated with the  $\theta$ -th quantile and  $\varepsilon_{i\theta}$   
 183 is an unknown error term. According to Koenker and Bassett (1978) who introduced QRMs the  $\theta$ -th  
 184 regression quantile,  $0 < \theta < 1$ , is defined as any solution to the minimization of the sum of absolute  
 185 deviation residuals:

$$\min_{\boldsymbol{\beta} \in R^k} \left\{ \sum_{i: y_i \geq \mathbf{x}_i' \boldsymbol{\beta}} \theta |y_i - \mathbf{x}_i' \boldsymbol{\beta}| + \sum_{i: y_i < \mathbf{x}_i' \boldsymbol{\beta}} (1 - \theta) |y_i - \mathbf{x}_i' \boldsymbol{\beta}| \right\} \quad (2)$$

187 which is solved by linear programming methods. When  $\theta$  is continuously increased from 0 to 1, we  
 188 obtain the entire conditional distribution of  $y$  conditional on  $\mathbf{x}$ . Starting from the general equation (1)  
 190 and with the aim of identifying factors associated with values of the two measures estimated, two  
 191 quantile regression models which assumed the dependent variable  $y_i$  ( $i=1, \dots, n$ ) to be: (i) H'EA and  
 192 (ii) H'MR respectively were estimated. In both models the  $k$ -dimensional vector  $\mathbf{x}_i$  of covariates  
 193 includes factors describing the land use category, pH and the level of TOC. Among the soil  
 194 properties that mostly affect microbial biomass activity and diversity, TOC and pH were chosen as  
 195 covariates to explain H' index variability (Fierer and Jackson, 2006; Constancias et al. 2015). In this

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196 study, the distribution of TOC values allowed to homogeneously group soils into three categories: i)  
197 low: for TOC < 1.5 %; ii) medium TOC < 1.5 – 3>; iii) high: for TOC  $\geq 3\%$ . Similarly, pH values  
198 allowed to group soils into three categories: i) <6.5 slightly acid – very strongly acid, ii) <6.5-7.4>  
199 neutral, iii) >7.4 slightly alkaline – moderately alkaline. The grouping criteria were established with  
200 the aim to obtain three groups with the same number of observations.

201 STATA software (STATA 13.2 edition) was used for statistical analyses. Three different levels of  
202 significance were considered for the estimated coefficients and are reported in the model: a value of  
203  $p < 0.01$  (indicated in the tables of results with \*\*\*), emphasizing strong relationships between the  
204 explicative variable of interest and the dependent variable; the value of  $p < 0.05$  (indicated in tables of  
205 results with \*\*) and finally a value of  $p < 0.10$  (indicated in the tables of results with \*) emphasizing a  
206 weak relationship between the variables.

207

## 208 Results

209 Figure 1 shows the functional capacity of soil microbial biomass calculated as the SEI and SIR for all  
210 soils within the three categories (F, A and EC). Extreme conditions soils showed the highest level of  
211 variability, including upper and lower outliers, for SEI (ranging from 11821 to 42 nmoles MUF  $g^{-1} h^{-1}$ ),  
212 while forest soils functional capacity was more variable for SIR (ranging from 177 to 0.9  $\mu g$   
213  $CO_2 g^{-1} h^{-1}$ ). Agricultural soils show, for both methodological approaches, a smaller range of  
214 variation and lower outliers.

215 Figure 2 shows the distribution of the two indexes values H'EA and H'MR, respectively, over the  
216 196 values of microbial functional diversity.

217 The two distributions are positively skewed and leptokurtic - as emerged by the descriptive statistics  
218 reported in Table 2 – and significantly different from normality as confirmed by the Shapiro-Wilk W  
219 and Shapiro-Francia W' test (the null hypothesis of normal data are rejected for both distributions, p-  
220 value=0.000).

221 A similar level of overall variability (which also includes upper and lower outliers) characterizes the  
222 two distributions as described by the values of coefficient of variation (CV). On the other hand, by  
223 focusing on the box-plots in Figure 3, it should be noted that – while considering the different  
224 magnitude of the two indexes – the larger height of the rectangles highlights a greater level of  
225 variability in the middle part of the distribution (i.e. the central half of the sample) concerning H'MR  
226 index. The presence of values outside the whiskers (dots in Figure 3) identifies lower outliers for  
227 H'EA and upper outliers for H'MR index.

228 The Spearman rank correlation, verifying the similarity of the orderings of the data when ranked  
229 according to each of the measures, showed that the two measures are not related for measuring  
230 microbial functional diversity ( $r_s = -0.0355$ ; p-value = 0.6217) (Table 3).

231 However, by distinguishing rank correlation according to the land use category, we found a moderate  
232 level of concordance ( $r_s = 0.2213$ ; p-value = 0.0656) when the two indexes refer to soil of type A. No  
233 correlation was found between the two measures for soil type EC and F ( $r_s = -0.1410$  and  $r_s = -$   
234  $0.0579$  respectively) (Table 3). According to the results of Kruskal-Wallis test, both H'EA and H'MR  
235 significantly distinguished the various soils when TOC or pH ranges were considered (Figure 4).  
236 However, H'MR showed a greater effectiveness than H'EA according to the p-values reported in  
237 Figure 3. In fact, while H'EA discriminated soils only for TOC values <1.5 - 3%> and pH values  
238 <6.5 - 7.4>, H'MR was significantly effective along all ranges for both soil properties.

239 The analysis of the potential relationships between each of the measures (H'EA and H'MR) and the  
240 two selected variables (TOC and pH) was carried out by referring to the quantile regression  
241 approach, which enabled us to analyse the effect of the covariates (TOC and pH) throughout the  
242 entire distribution as well as at the extremes. Table 4 shows the estimation results of regression  
243 models at quantiles 0.25, 0.50 and 0.75.

244 The total organic carbon content is not significantly associated neither with H'MR nor with H'EA,  
245 except for medium and high classes of TOC values at quantile 0.50 for H'EA (p-value<0.10).  
246 Conversely, pH values are negatively related to the H'EA measure in the lower part of the

247 distribution (i.e. for low values of the dependent variable H'EA) while a positive relationship was  
248 observed in the highest quantiles of the distribution (i.e. for high values of the dependent variable  
249 H'EA). On the other hand, pH values are positively related with H'MR in the middle part of the  
250 distribution (quantile 0.50) only.

251 The land use category is an important factor distinguishing the values of the two measures. For H'EA  
252 the relationship is positive and strongly significant in the lower part of the distribution (quantile 0.25)  
253 while a negative relationship in the upper part (quantile 0.75) was found. A negative relationship  
254 characterizes this land use category (A) and the H'MR measure in the first half part of the  
255 distribution (quantiles 0.25 and 0.50) compared to the EC category representing the reference  
256 category. Finally F soils category shows a positive relationship with H'MR only in the lower part of  
257 the distribution (quantile 0.25)(Table 4).

258

## 259 Discussion

260 In this study, a large data set of 196 values of Shannon diversity index, calculated from data of  
261 enzyme activities and CLPP-MicroResp techniques, was used. Griffiths et al. (2016) recently  
262 included both techniques in a list of 18 potential, powerful indicators aimed to monitor soil  
263 biodiversity and ecosystem function across Europe.

264 The first aim of this paper was to assess the relative sensitivity of each methodological approach in  
265 capturing differences among the land use categories when different levels of pH and TOC are  
266 considered.

267 The Kruskal-Wallis test showed that CLPP-MicroResp was a more powerful technique than enzyme  
268 activities in highlighting differences among land use categories. Remarkably, while enzyme activities  
269 were effective only within a certain range of TOC and pH values (<1.5-3%> and <6.5-7.4>,  
270 respectively), MicroResp was able to discriminate soils along the whole range of TOC and pH  
271 values, thus representing an effective tool for evaluating microbial functional diversity changes. This  
272 result might be explained by the fact that differences in soil microbial catabolic evenness among

273 various land-uses are usually related to differences in organic C pools (Degens et al., 2000).

274 Moreover, similar results were found by Creamer et al. (2009) reporting that the MSIR (multi  
275 substrate induced respiration) technique resulted in a much more distinct and relatively consistent  
276 pattern of separation between the tested soils with respect to enzyme activities.

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277 However, in relation to the lack of significant response of EA to pH variations we should keep in  
278 mind that enzymes determination requires NaAc buffer pH 5.5 as standardized in the protocol  
279 proposed by Marx et al. (2001). It is thus possible that the lower discriminant capacity of enzyme  
280 activities across a wide range of pH values may be ascribable to this methodological constraint.  
281 Nevertheless, since also TOC values did not affect significantly H'EA, except in the range <1.5-  
282 3.0%>, we can conclude that MicroResp showed a higher discrimination capacity among soil uses  
283 and managements.

284 In this study, no correlation was found between H'EA and H'MR over all the data collected.

285 However, when looking at the correlation between H'EA and H'MR within the three categories of  
286 soils (A, F and EC) a weak relationship (significant at 10% level) emerges only in arable soils. We  
287 cannot exclude that this is due to the fact that agricultural soils features mainly fall within the  
288 intermediate range of pH and TOC values where both methods are sensitive.

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289 The general lack of correlation between enzymes and CLPP-MicroResp confirmed that the two  
290 techniques assess different steps of decomposition processes. Enzymatic hydrolysis focuses on the  
291 breakdown of complex organic polymers, which not necessarily leads to the complete mineralization  
292 of substrates but can also lead to anabolic pathways for biosynthetic processes, polymerization,  
293 condensation (i.e. humification, interaction with mineral colloids). Conversely, CLPP-MicroResp  
294 measures the complete mineralization of simple and complex organic compounds to CO<sub>2</sub>, which  
295 represents the final step of decomposition process. Therefore, in our opinion, a comprehensive  
296 assessment of microbial functional diversity can be provided by the integration of both techniques.  
297 For this reason, they can be considered complementary components of microbial functional diversity  
298 providing a different ecological significance. In particular, in a comparison within soils we might

299 suppose that if H'EA:H'MR ratio decreases, the diversity of soil functions is completely oriented  
300 towards mineralization of organic matter. In this case, it can be evidenced a dissipating energy  
301 system; however this last hypothesis, aimed to assign a different ecological significance to the two  
302 methods, needs to be further deepened and confirmed. Regarding the ecological significance of  
303 increased catabolic diversity due to land use, it might be also supposed an amplified resistance of  
304 microbial communities to stress or disturbance (Degens et al. 2001). Moreover, to further explain the  
305 lack of correlation between the two methods, we should keep in mind that soil enzymes include the  
306 contribution, considerable in most cases, of the immobilized fraction (humus-clay bound enzymes)  
307 (Nannipieri et al. 2012). This fraction is considered a permanent bio-catalytic property of the soil, not  
308 necessarily linked to the living biomass. Immobilized enzymes may represent soils background  
309 hydrolytic potential, established and stabilized during time, and representing their resilient capacity  
310 (Ceccanti et al. 2008). To date, no methods are available to distinguish between the extracellular  
311 activity of stabilized enzymes from that of enzymes associated with active cells. Such separation is  
312 important because only enzymes associated with active cells contribute to microbial activity. The  
313 stabilized extracellular fraction is no more related to microbial metabolism and can persist in soil  
314 under unfavourable conditions for soil microorganisms (Nannipieri et al. 2012).

315 Therefore, enzyme activities, and the functional diversity measures derived from using this  
316 methodology, inform on the general soil biological functioning including not only the actual living  
317 microbial activity but also the past biochemical activity still operating within soil matrix. Conversely,  
318 CLPP-MicroResp has been considered a direct measurement of microbial communities' catabolic  
319 profile providing an instant photograph of microbial physiology (Lagomarsino et al. 2007).

320 The QRM helped to understand if, and to what extent, the role of selected covariates (relevant soil  
321 properties such as TOC and pH) change throughout the entire distribution of each dependent variable  
322 (H'EA and H'MR). The QRM showed that both diversity indexes depended more on soil pH than on  
323 total organic carbon content indicating soil reaction as the property mostly affecting microbial  
324 diversity (Zahlina et al. 2014).

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325 Microbial functional diversity expresses the capacity of microbial community to perform different  
326 processes and to metabolize diverse substrates. Soil pH variations can induce, more than the mere  
327 TOC content, significant changes within microbial biomass structure in terms of species and related  
328 functional patterns. Microbial biochemical processes are strictly dependent on pH values that control  
329 the majority of the reactions occurring in the soil. Fierer et al. (2006) and Lemanceau et al. (2015)  
330 reported soil pH as the best predictor of microbial diversity and richness affecting consequently  
331 microbial functions. However, the nature of this relationship is controversial. Griffiths et al. (2011)  
332 report that no decline in diversity was observed at increasing pH in a spatial assessment of soil  
333 bacterial community profiles across Great Britain. Fierer et al. (2006), in a similar study performed  
334 across North and South America, showed a unimodal distribution of bacterial diversity, reaching  
335 possibly a plateau at near neutral pH.

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336

## 337 CONCLUSIONS AND FUTURE PERSPECTIVES

338 This study proved that CLPP-MicroResp technique provided a higher discrimination capacity, if  
339 compared to enzyme activities, as an ecological indicator to assess soil microbial functional diversity.  
340 In relation to soil chemical properties, pH was more relevant than TOC content in differentiating  
341 processes carried out by microorganisms. The diversity indexes obtained by the two methods, EA  
342 and MR, were not correlated; we hypothesize that they target complementary components of  
343 microbial functional diversity. This study could be improved in the future with the aim to verify if the  
344 two methodological approaches provide a different ecological significance informing on the extent of  
345 dissipating energy pathways in the soil system.

346

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## Highlights

- ✓ Enzymes and MicroResp as reliable indicators to assess microbial functional diversity
- ✓ No correlation was found between the enzyme and MicroResp diversity indexes
- ✓ The two methods target complementary components of microbial functional diversity
- ✓ Both methods were effective to show differences among various land use categories
- ✓ Quantile regression model allowed analysis along the distribution diversity indexes

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4 1 **ASSESSMENT OF SOIL MICROBIAL FUNCTIONAL DIVERSITY: LAND USE AND**  
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6 2 **SOIL PROPERTIES AFFECT CLPP-MICRORESP AND ENZYMES RESPONSES**  
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10 4 **Moscatelli M.C.<sup>1</sup>, Secondi L.<sup>1</sup>, Marabottini R.<sup>1</sup>, Papp R.<sup>1</sup>, Stazi S.R.<sup>1</sup>, Mania E.<sup>2</sup>, Marinari S.\*<sup>1</sup>**

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12 5 *<sup>1</sup>Dept. for Innovation in Biological, Agrofood and Forest systems (DIBAF), University of Tuscia,*  
13  
14 6 *Viterbo, Italy*

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16 7 *<sup>2</sup>Department of Agricultural and Forest Sciences (DISAFA), University of Torino, Torino, Italy*  
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30 13 **Keywords: microbial processes, Shannon index, Soil properties.**  
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63 **14 Abstract**  
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65 15 The assessment of microbial functional diversity is an important indicator of soil quality. Different  
66 16 methodological approaches are currently used; among them are enzyme activities (EA) and CLPP  
67 17 (community level physiological profile) techniques (e.g. MicroResp<sup>TM</sup>, MR). The aims of the study  
68 18 were: *i*) to assess the efficacy of both methods in capturing differences among various land use  
69 19 categories when different levels of selected explanatory variables such as land use category, total  
70 20 organic carbon (TOC) and pH are considered, and *ii*) to explore, through a quantile regression  
71 21 approach, the possible relationships between each of the two methods with land use category, TOC  
72 22 and pH. The Shannon diversity index ( $H'$ ), calculated from EA and MR data, was chosen as a  
73 23 synthetic index deriving from the same mathematical model. The quantile regression model (QRM),  
74 24 the Kruskal-Wallis and Spearman rank correlation tests were performed.

75 25 Enzyme activities and MicroResp were reliable ecological indicators to assess soil microbial  
76 26 functional diversity. No correlation was found between the diversity indexes,  $H'$ EA and  $H'$ MR, it  
77 27 was therefore supposed that the two methods may target complementary components of microbial  
78 28 functional diversity. Both methods were effective in capturing differences among various land use  
79 29 categories, in particular  $H'$ MR in soils with low TOC content (<1.5%). Moreover, the QRM  
80 30 approach allowed a more detailed analysis along the distribution of the diversity indexes ( $H'$ EA and  
81 31  $H'$ MR) indicating that  $H'$ EA was more dependent on the selected variables.

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## 34 **Introduction**

35 The links between ecosystem functioning and levels of soil biodiversity have been the focus of the  
36 recent scientific literature (Delgado-Baquerizo et al., 2016; Creamer et al., 2016b; Griffiths et al.,  
37 2016; Nannipieri et al., 2003). The first authors provided evidence that loss in microbial diversity  
38 will likely reduce multiple ecosystems functions thus negatively impacting the provision of  
39 ecosystem services. Adhikari & Hartemink (2016) claimed for new insights into soil microbial  
40 diversity and their role in soil functional variability. Since up to 80/90% of soil functions, from  
41 humification to mineralization, is microbially-mediated, the diversification of soil microorganisms in  
42 terms of structure and/or activity is essential to maintain functioning of terrestrial ecosystems (Pereira  
43 et al., 2013).

44 Microbial functional diversity is defined as “the sum of the ecological processes, and/or capacity to  
45 use different substrates developed by the organisms of a community” (Insam et al., 1989). Emmerling  
46 et al. (2002) and Wellington et al. (2003) report that if microbial genetic diversity assesses a latent  
47 diversity, which may not be expressed, functional diversity is related to the actual activities resulting  
48 from that potential so that "functional rather than taxonomic diversity may provide greater insight to  
49 microbial roles in ecosystems" (Zak et al., 1994).

50 Over the last 10 years, the scientific literature provided a great number of papers aimed to assess  
51 microbial functional diversity as an important ecological indicator to monitor and assess soil quality  
52 changes in different pedoclimatic conditions, land uses and human pressure levels (e.g. management  
53 practices) (Bardgett and van der Putten, 2014; Griffiths et al., 2016).

54 To measure the activity and diversity of the microbial community a number of methods can be  
55 applied, to cite few of the most common approaches: (i) catabolic activity investigated by Biolog<sup>TM</sup>-  
56 plates (Garland and Mills, 1991; Rutgers et al., 2016), (ii) respiration of different substrates as  
57 investigated by the MicroResp<sup>TM</sup> method (Campbell et al., 2003; Chapman et al., 2007; Creamer et  
58 al. 2016a) and (iii) enzyme activities (Nannipieri et al., 2012; Hendriksen et al., 2016).

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183 59 Although all methodological approaches are reliable and sensitive, few studies aimed to understand  
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185 60 their effectiveness to discriminate microbial functional diversity in relation to soil organic C and pH  
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187 61 as the main properties being affected by land use and management practices, anthropic impact and  
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190 62 other pedogenic factors. To achieve this goal, a large number of case studies covering different land  
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192 63 use categories is necessary. In this study, about 200 measurements of microbial functional diversity  
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194 64 obtained over a broad spectrum of key soil properties and across different land uses and management,  
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196 65 were selected. Furthermore, microbial functional diversity obtained through enzyme activities (EA)  
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198 66 and CLPP-MicroResp (MR), was synthetically represented by the Shannon index ( $H'$ ) that  
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200 67 transforms the obtained results to a comparable range of values deriving from the same mathematical  
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202 68 model. The Shannon index is a comprehensive indicator of microbial species, individual numbers and  
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204 69 evenness, or distribution of the enzyme activities and is influenced by richness of community species  
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207 70 (Bending et al., 2002; Li et al., 2007).

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209 71 The aim of the present study was therefore to: i) assess the efficacy of both methods in capturing  
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211 72 differences among the different land use categories when different levels of pH and TOC are  
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213 73 considered, ii) explore, through a quantile regression approach, the possible relationships between  
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215 74 each of the two methods and selected explanatory variables (TOC, land use category, pH).  
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## 219 76 **Materials and methods**

### 220 221 77 *Experimental design, sites and soil categories*

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223 78 The results presented in this paper have been obtained performing additional statistical analyses on  
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226 79 data collected in the Laboratory of Chemistry and Biochemistry, University of Tuscia, Viterbo, Italy  
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228 80 during the last 6 years (2010-2016). Microbial functional diversity was measured, by means of  
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230 81 enzyme activities and CLPP-MicroResp<sup>TM</sup> technique, in a wide range of soils analysed within  
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232 82 different research projects. Most of the sampling sites are located within the Mediterranean climatic  
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234 83 area. Other climatic areas are the monsoon one for the Bangladesh case study, the temperate one for  
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236 84 Switzerland, oceanic for United Kingdom and boreal for Sweden. All soils represent a broad  
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85 spectrum of key soil properties across different land use categories, wide range of soil pH and soil  
86 organic carbon content (TOC) (Table 1).

87 The soils were related to 15 case studies, each one including different treatments, with the aim to  
88 separate diverse land uses and/or specific conditions. For this purpose, three groups were identified:  
89 F (forest soils, 4 case studies), A (agricultural soils, 5 case studies) and EC (extreme conditions, 6  
90 case studies). The case studies related to forest soils (F) included different management practices,  
91 lithological substrates, afforestation and chronosequences. The soils under agricultural land use (A)  
92 were characterized by different managements and/or agricultural practices such as: organic,  
93 biodynamic, conventional cropping systems, tillage/no tillage and natural green cover/no cover. The  
94 third category (EC) included soils with peculiar characteristics due to either pedo-climatic conditions  
95 (saline environments, natural arsenic contamination in rice paddies and highly calcareous soils) or to  
96 heavy anthropic impact (thallium contamination, a multi-element contaminated dump, arsenic  
97 contaminated mine) (Table 1).

### 99 *Soil sampling*

100 All soils were sampled at 0-20 cm depth during the dry season (spring/summer), air dried, sieved at 2  
101 mesh and preserved at room temperature., Then, prior to biochemical analyses, soil moisture content  
102 of air dried samples was adjusted to 60 % of their water holding capacity and soils were re-  
103 conditioned for 10 days .

### 105 *Soil analyses and methodologies*

106 The total organic carbon (TOC) was determined by combustion by Shimadzu TOC VCSH analyzer  
107 while soil pH was measured on sieved soil suspended in a solution of deionised water in 1:2.5 ratio  
108 (w/v). The pH was measured in the supernatant with a pH meter (pH 211, Hanna Instruments).

109 A total of 196 values of microbial functional diversity, assessed by means of enzyme activities and  
110 CLPP-MicroResp, were used for this study (Table 1). Enzymes were measured following Marx et al.

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303 111 (2001) using fluorogenic methylumbelliferyl (MUF)-substrates. Soils were analysed for  
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306 112 cellobiohydrolase,  $\beta$ -1,4-glucosidase,  $\alpha$ -1,4-glucosidase,  $\beta$ -N-acetyl-glucosaminidase,  $\beta$ -1,4-  
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308 113 xylosidase, acid-phosphatase, arylsulphatases and butyrate esterase which is considered a proxy of  
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310 114 endocellular activity (Wittman et al. 2004). The relative fluorogenic substrates, prepared with acetate  
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312 115 buffer 0.5 M pH 5.5, were: 4-MUF- $\beta$ -D-cellobioside, 4-MUF- $\beta$ -D-glucoside, 4-MUF-N-acetyl- $\beta$ -  
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314 116 glucosaminide, 4-MUF- $\alpha$ -D-glucoside, 4-MUF-phosphate, 4-MUF-7- $\beta$ -D-xyloside, 4-MUF-sulphate  
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316 117 and 4-MUF-butyrate. Fluorescence (excitation 360 nm, emission 450 nm) was measured with an  
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318 118 automatic fluorimetric plate-reader (Fluoroskan Ascent) and readings were performed after 0, 30, 60,  
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320 119 120 and 180 minutes of incubation at 30° C. The results were expressed as nmoles of product (MUF)  
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323 120 of each enzymatic reaction released per g of soil per unit of time in relation to a standard curve  
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325 121 prepared with increasing MUF concentrations and incubated at the same experimental conditions.

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327 122 The community level physiological profile (CLPP) was determined using the MicroResp™ soil  
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329 123 respiration system (James Hutton Ltd, Aberdeen, UK) according to Campbell et al. (2003).

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331 124 The 15 substrates used for MicroResp were:  $\alpha$ -D-glucose, D-Galactose, D-fructose, L-arabinose, L-  
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333 125 leucine, L-arginine, Glycine, L-aspartic acid,  $\gamma$ -amino-butyric and glutamic acid, three carboxylic  
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335 126 acids: citric acid, oxalic acid and L-ascorbic acid, and two phenolic acids: vanillic and syringic acid.

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337 127 The emission of CO<sub>2</sub> by the microbial biomass was estimated using a colorimetric method  
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340 128 (microplate spectrophotometer) before and after 6 h of incubation at 28 °C. The absorbance was read  
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342 129 at 595nm. At the end the absorbance was normalised for any difference recorded at time zero and  
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344 130 then converted to % CO<sub>2</sub> using the calibration curve  $y = A+B/(1+ D \times Ai)$  (Campbell et al., 2003).

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346 131 The calibration procedure was performed taking into account the spectrophotometer used, the  
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348 132 different soils and incubation conditions. In our experimental conditions the constants of the equation  
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350 133 were: A:-1,62, B:-4,85 and D: -8,1. The CO<sub>2</sub>% was converted to  $\mu\text{g C-CO}_2 \text{ g}^{-1} \text{ h}^{-1}$  production rate  
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352 134 using gas constant, T °C, headspace volume, soil dry weight (d.w.) and incubation time. The SEI  
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354 135 (Synthetic Enzymatic Index) and SIR (Substrate Induced Respiration) for all soils within the three  
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356 136 categories (F, A and EC) have been calculated as synthetic measures of microbial functional  
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363 137 capacity. Both SEI and SIR represent the total microbial functional capacity expressed as sum of all  
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365 138 enzymatic activities and of induced respiration of all substrates, respectively.  
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368 139 Microbial functional diversity was assessed calculating the Shannon-Weaver diversity index  
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370 140 (Kennedy and Smith, 1995) corresponding to the entropy concept defined by:  $H' = - \sum p_i * \ln p_i$   
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372 141 (Shannon and Weaver, 1949; Spellerberg and Fedor, 2003), where  $p_i$  is in turn: for H'EA, the ratio  
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374 142 of the activity of a particular enzyme to the sum of all enzymatic activities while for H'MR it is the  
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376 143 ratio of the respiration rate of each single C-substrate to the sum of all substrates. Shannon diversity  
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378 144 index is related to the entropy of a system and when applied as a measure of microbial functions  
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380 145 entropy, may express the heterogeneity of soil organic substrates availability and microbial processes  
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382 146 (Marinari et al., 2013). Since the eight enzymes and the 15 substrates here tested did show activity in  
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384 147 all the analysed samples, then, in this work, the diversity recorded reflects only the “evenness” or  
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387 148 distribution of the enzyme activities or ability to use the different substrates (Bending et al., 2002;  
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389 149 Rodríguez-Loinaz et al., 2008).

### 391 150 392 393 151 *Statistical analyses*

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395 152 The analysis of all collected data was carried out into various steps. At first, with the aim to compare  
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397 153 the two indexes, a standardization due to the existing differences in the range of H'EA and H'MR  
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399 154 possible/admissible values was performed. As usual the new standardized indexes have mean equal  
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401 155 to zero and variance equal to 1. It is worth noting that from now on, all the statistical analyses were  
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404 156 carried out on the two standardized distributions.

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406 157 The descriptive analyses provided a clear picture of the distribution of the two indexes (H'EA and  
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408 158 H'MR) as well as information about the shape of the two distributions. Moreover, rank correlation  
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410 159 measures and test performed by using the Spearman correlation enabled to evaluate if, and to what  
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412 160 extent, the two methodological approaches (EA, enzyme activities and MR, MicroResp™) used to  
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414 161 evaluate soil microbial functional diversity are related.

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423 162 The Kruskal-Wallis non-parametric test was used to test if and to what extent the two indexes  
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425 163 distinguished the various land use categories in relation to TOC and pH ranges.  
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428 164 By considering the asymmetry of the two distributions (e.g. H'EA and H'MR respectively, as shown  
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430 165 in Figure 2) as well as the results of the Shapiro-Wilk normality test, we analysed the existence of  
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432 166 association between each of the two measures and selected covariates by using Quantile Regression  
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434 167 Models (QRMs). In fact, these types of regression models offer the possibility to highlight how the  
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436 168 effect of the selected covariates, in this case TOC content, pH and land use category, changes  
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438 169 throughout the entire distribution of the dependent variable. To estimate the relationships (in terms of  
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440 170 association) between the dependent variables and the set of selected covariates, the classical OLS  
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442 171 (Ordinary Least Squares) regressions could be applied. However, data obtained from experimental  
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444 172 collection tend to be skewed so that these models could not be able to describe the "correct"  
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446 173 relationships. Moreover, QRMs are more robust to the presence of outliers and can be consistent  
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448 174 under weaker stochastic assumption than with least-squares estimation (Cameron and Trivedi, 2005;  
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450 175 Koenker, 2005).  
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453 176 Referring to the soil context, the application of QRMs has important advantages. Firstly, QRMs can  
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455 177 help to explore if the existing differences observed between the two measures can be attributed to  
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457 178 different effects played by the explanatory variables at the various quantiles. Secondly, it can be  
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459 179 interesting to understand what happens throughout the entire distribution of the two measures (H'EA  
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461 180 and H'MR) and at their extremes.  
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464 181 We estimated two QRMs which assumed the dependent variable to be: (i) H'EA and (ii) H'MR  
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466 182 respectively. In both models the set of covariates includes factors describing: the land use category  
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468 183 (distinguished into Forest, Agricultural and Extreme soil Conditions), the levels of pH and TOC.  
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470 184 Among the soil properties that mostly affect microbial biomass activity and diversity, TOC and pH  
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472 185 were chosen as covariates to explain H' index variability (Creamer et al., 2016b; Fierer and Jackson,  
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474 186 2006; Constancias et al. 2015). In this study, the distribution of TOC values allowed to  
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476 187 homogeneously group soils into three categories: i) low: for TOC < 1.5 %; ii) medium TOC < 1.5 –  
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483 188 3>; iii) high: for TOC  $\geq 3\%$ . Similarly, pH values allowed to group soils into three categories: i) <6.5  
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485 189 slightly acid – very strongly acid, ii) <6.5-7.4> neutral, iii) >7.4 slightly alkaline – moderately  
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488 190 alkaline. STATA software (STATA 13.2 edition) was used for statistical analyses. Three distinct  
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490 191 levels of significance were considered for the estimated coefficients and are reported in the model: a  
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492 192 value of  $p < 0.001$  (indicated in the tables of results with \*\*\*), emphasizing strong relationships  
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494 193 between the explicative variable of interest and the dependent variable significant at 0.1% level; the  
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496 194 value of  $p < 0.01$  (indicated in tables of results with \*\*) indicates a relationship significant at 1% level  
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498 195 and finally a value of  $p < 0.05$  (indicated in the tables of results with \*) emphasizing a relationship  
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500 196 between the variables significant at 5% level.  
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## 505 198 **Results**

507 199 Figure 1 shows the functional capacity of soil microbial biomass calculated as the SEI and SIR for all  
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510 200 soils within the three categories (F, A and EC). Soils characterized by extreme conditions showed  
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512 201 the highest level of variability - including upper outliers - for SEI (ranging from 42 to 11800 nmoles  
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514 202 MUF  $g^{-1} h^{-1}$ ), while the functional capacity of forest soils showed a high level of dispersion for SIR  
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516 203 (ranging from 0.9 to 177  $\mu g CO_2 g^{-1} h^{-1}$ ). Agricultural soils show, for both methodological  
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518 204 approaches, a smaller level of variability.

520 205 Figure 2 shows the distribution of the two standardized indexes values H'EA and H'MR,  
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522 206 respectively, over the 196 values of microbial functional diversity. The two distributions are  
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524 207 positively skewed and leptokurtic - as emerged by the descriptive statistics reported in Table 2 – and  
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527 208 significantly different from normality as confirmed by the Shapiro-Wilk W and Shapiro-Francia W'  
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529 209 test.

531 210 The Spearman rank correlation, verifying the similarity of the orderings of the data when ranked  
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533 211 according to each of the measures, showed that the two measures are not related for measuring  
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535 212 microbial functional diversity ( $p$ -value = 0.0987) (Table 3). However, by distinguishing rank



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543 213 correlation according to the land use category, we found a moderate and significant level of inverse  
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545 214 rank correlation (p-value = 0.0073) when the two indexes refer to soil of type A. No significant rank  
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548 215 correlation was found between the two measures for soil type EC and F (p-value= 0.6534 and p-  
549  
550 216 value= 0.8727 respectively) (Table 3).

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552 217 According to the results of Kruskal-Wallis test, both H'EA and H'MR distinguished in different ways  
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554 218 the various soils when TOC or pH ranges were considered (Table 4). Thus, according to the obtained  
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556 219 p-values H'MR showed a slightly higher discriminatory potential than H'EA. H'MR, in fact, was  
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558 220 significantly effective at low TOC ranges (<1.5%) where H'EA was not. On the other hand, both  
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560 221 methods failed to discriminate in alkaline soils (pH values  $\geq 7.4$ ).

562 222 The analysis of the potential relationships between each of the measures (H'EA and H'MR) and the  
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564 223 selected variables (land use, TOC and pH) was carried out by referring to the quantile regression  
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566 224 model (QRM), which enabled to analyse the effect of the covariates throughout the entire distribution  
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568 225 as well as at the extremes. Table 5 shows the estimation results of regression models at quantiles  
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571 226 0.25, 0.50 and 0.75.

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573 227 Focusing on TOC content we only found a negative association at quantile 0.75 (p-value<0.05)  
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575 228 between H'EA and high level of TOC (equal or greater than 3%). On the other hand, pH levels are  
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577 229 negatively related to the H'EA measure in the lower part of the distribution (e.g at quantile 0.25 of  
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579 230 the dependent variable H'EA, p value < 0.01) while a positive relationship was observed with high  
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581 231 levels of pH in the highest quantile of the distribution (e.g. for high values of the dependent variable  
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583 232 H'EA). Furthermore, a positive relationship was found between medium level of pH (values ranging  
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585 233 between 6.5 and 7.4) and H'MR in the middle part of the distribution (quantile 0.50).

587  
588 234 The land use category is a key factor distinguishing the values of the two measures. For H'EA the  
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590 235 relationship is positive and strongly significant at quantiles 0.25 and 0.75 for land use category F  
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592 236 (forest soils) while a negative relationship with agricultural land use category was observed at  
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594 237 quantiles 0.50 and 0.75. At the same time, we observed positive and significant relationships between  
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603 238 the values of H'MR and agricultural land use category at all different quantiles throughout the entire  
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605 239 distribution while forest soil only at 0.75 quantile (Table 5).  
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## 609 610 241 **Discussion**

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612 242 In this study, a large data set of 196 values of Shannon diversity index, calculated from data of  
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614 243 enzyme activities and CLPP-MicroResp techniques, was used. Griffiths et al. (2016) recently  
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616 244 included both techniques in a list of 18 potential, powerful indicators aimed to monitor soil  
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618 245 biodiversity and ecosystem function across Europe.

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620 246 The first aim of this paper was to assess the relative sensitivity of each methodological approach in  
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622 247 capturing differences among the land use categories when different levels of pH and TOC are  
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625 248 considered.

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627 249 The Kruskal-Wallis test showed that both methods were able to highlight differences among land use  
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629 250 categories at almost all ranges of TOC and pH. However, while both of them failed to discriminate in  
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631 251 alkaline soils (pH >7.4), only MicroResp was completely effective along the whole TOC gradient,  
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633 252 including low TOC values (<1.5%). This result might point to MicroResp as a more powerful tool for  
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635 253 evaluating microbial functional diversity, particularly in oligotrophic environments where the  
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637 254 addition of easily available organic C sources (represented by the different substrates) may stimulate  
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639 255 microbial respiration. Conversely, enzyme production is not similarly stimulated as it requires a  
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641 256 higher energetic expense (Burns and Dick, 2002). In studies aimed to evaluate the effect of land use  
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644 257 change on microbial functional diversity, the CLPP-MicroResp approach can be thus suggested as  
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646 258 soil microbial catabolic evenness among various land-uses is usually related to differences in organic  
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648 259 C pools (Degens et al., 2000). Creamer et al. (2009) also reported that the MSIR (multi substrate  
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650 260 induced respiration) technique resulted in a much more distinct and relatively consistent pattern of  
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652 261 separation between the tested soils with respect to enzyme activities. The lack of potential for both  
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654 262 techniques to discriminate among different land uses in alkaline soils (pH >7.4) may be due to the  
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663 263 fact that the interrelationship between soil pH and microbial diversity may be lost (Fierer et al., 2006)  
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665 264 or even decreased (Griffiths et al., 2011) at soil pH values higher than 7.  
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668 265 In this study, no correlation was found between H'EA and H'MR all over the data collected.  
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670 266 Moreover, an opposite behaviour of the two indexes was found in agricultural soils where the  
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672 267 significant ( $p < 0.01$ ) correlation coefficient was negative. This result confirms what was previously  
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674 268 observed regarding oligotrophic environments characterized by lower organic matter content, such as  
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676 269 agricultural soils. In fact, as reported by Lagomarsino et al. (2011), the microbial functional diversity  
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678 270 determined by means of the enzymatic pattern is affected by land use showing an increase along a  
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680 271 gradient of soil organic matter. In the same paper the authors reported an inverse relationship  
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682 272 between microbial functional diversity and the catabolic response per unit of biomass expressed by  
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684 273 the metabolic quotient ( $qCO_2$ ).  
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687 274 The lack of correlation between H' by means of enzymes and CLPP-MicroResp suggests that the two  
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689 275 techniques may assess sequential steps of decomposition processes, even if in this meta-analysis the  
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691 276 product of most selected enzymatic reactions did not represent the substrates used to test CLPP-  
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693 277 MicroResp. Enzymatic hydrolysis focuses on the breakdown of complex organic polymers, which  
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695 278 not necessarily leads to the complete mineralization of substrates but can also lead to anabolic  
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697 279 pathways for biosynthetic processes, polymerization, condensation (e.g. humification, interaction  
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699 280 with mineral colloids). Conversely, CLPP-MicroResp measures the complete mineralization of  
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702 281 simple and complex organic compounds to  $CO_2$ , which represents the final step of decomposition  
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704 282 process. Therefore, in our opinion, a comprehensive assessment of microbial functional diversity can  
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706 283 be provided by the integration of both techniques. For this reason, they can be considered  
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708 284 complementary components of microbial functional diversity.  
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710 285 Moreover, to further explain the lack of correlation between the two methods, we should keep in  
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712 286 mind that soil enzymes include the contribution, considerable in most cases, of the immobilized  
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714 287 fraction (humus-clay bound enzymes) (Nannipieri et al., 2012). This fraction is considered a  
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716 288 permanent bio-catalytic property of the soil, not necessarily linked to the living biomass.  
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723 289 Immobilized enzymes may represent soils background hydrolytic potential, established and stabilized  
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725 290 during time, and representing their resilient capacity (Ceccanti et al., 2008). To date, no methods are  
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728 291 available to distinguish between the extracellular activities of stabilized enzymes from that of  
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730 292 enzymes associated with active cells. Such separation is important because only enzymes associated  
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732 293 with active cells contribute to microbial activity. The stabilized extracellular fraction is no more  
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734 294 related to microbial metabolism and can persist in soil under unfavourable conditions for soil  
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736 295 microorganisms (Nannipieri et al., 2012). Therefore, enzyme activities, and the functional diversity  
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738 296 measures derived from using this methodology, inform on the general soil biological functioning  
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740 297 including not only the actual living microbial activity but also the past biochemical activity still  
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742 298 operating within soil matrix. Conversely, CLPP-MicroResp has been considered a direct  
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744 299 measurement of microbial communities' catabolic profile providing an instant photograph of  
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746 300 microbial physiology (Lagomarsino et al., 2007).  
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749 301 The QRM helped to understand if, and to what extent, the role of selected covariates (land use, TOC  
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751 302 and pH) changes throughout the entire distribution of each dependent variable (H'EA and H'MR). It  
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753 303 is, in fact, known that microbial functions are largely dependent on organic substrates availability and  
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755 304 soil reaction (Bardgett and van der Putten, 2017). The QRM was found to be an effective statistical  
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757 305 approach to analyse microbial functional diversity response in relation to the selected covariates,  
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759 306 particularly at the lowest (0.25) and highest (0.75) quantiles.  
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761 307 In this study QRM showed that both diversity indexes depended more on soil pH than on TOC  
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763 308 content indicating soil reaction as the property mostly affecting microbial diversity (Zhalnina et al.,  
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765 309 2014). In fact, only when TOC values were above 3% the H'EA was negatively affected suggesting  
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767 310 that the increase of soil available organic compounds may cause a negative feedback on microbial  
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769 311 hydrolytic reactions. On the contrary, it was more evident the relationship between pH and both  
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771 312 indices. H'EA was negatively related to pH in the 0.25 quantile indicating that low levels of this  
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773 313 index are more sensitive to soil pH variations (Griffiths et al., 2011). Conversely the dependence of  
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775 314 both indexes (H' MR and H'EA) from pH was positive at 0.50, and 075 quantiles, respectively. Soil  
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783 315 pH variations can induce, more than the mere TOC content, significant changes within microbial  
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785 316 biomass structure in terms of species and related functional patterns. Microbial biochemical  
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788 317 processes are strictly dependent on pH values that control the majority of the reactions occurring in  
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790 318 the soil. Fierer et al. (2006) and Lemanceau et al. (2015) reported soil pH as the best predictor of  
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792 319 microbial diversity and richness affecting consequently microbial functions.

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794 320 However, the nature of this relationship is controversial. Griffiths et al. (2011) report that a decline of  
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796 321  $\beta$ -diversity was observed at increasing pH in a spatial assessment of soil bacterial community profiles  
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798 322 across Great Britain. Fierer et al. (2006), in a similar study performed across North and South  
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800 323 America, showed a unimodal distribution of bacterial diversity, reaching possibly a plateau at near  
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803 324 neutral pH.

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805 325 Finally, the influence of the different land use categories was evident in some parts of the distribution  
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807 326 for both indexes, especially at 0.75 quantile. In particular, the effect of forest soils was always  
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809 327 positive, in most cases significant, for both indexes at all quantiles, confirming the strict relationship  
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811 328 existing between the forest environment and soil microbial diversity (Creamer et al., 2016b).

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## 815 330 **CONCLUSIONS AND FUTURE PERSPECTIVES**

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817 331 This study demonstrates that both methods, enzyme activities and MicroResp, are reliable ecological  
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819 332 indicators to assess soil microbial functional diversity. However, since no correlation was found  
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821 333 between the diversity indexes H'EA and H'MR, it was hypothesized that the two methods may target  
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823 334 complementary components of microbial functional diversity.

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826 335 The results lead to the following conclusions: *i*) both methods were effective in capturing differences  
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828 336 among various land use categories although MicroResp was more sensitive at low levels of soil  
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830 337 organic matter, *ii*) the QRM approach allowed a more detailed analysis along the distribution of the  
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832 338 diversity indexes (H'EA and H'MR) with H'EA showing a more significant dependence on the  
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834 339 selected variables.

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340 This study can lay the foundations to further studies aimed to assign an ecological significance to the  
341 assessment of microbial functional diversity.

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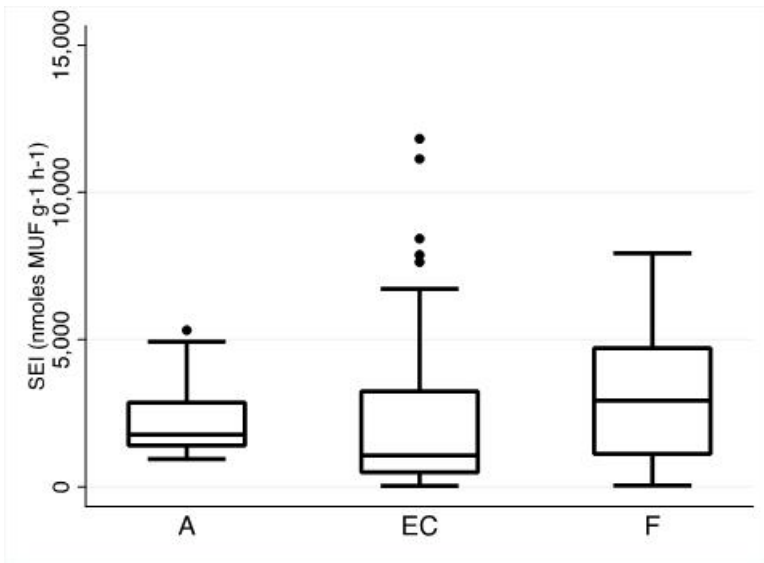
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## Figures

Figure 1: Boxplot of microbial functional capacity measured by means of enzyme activities and CLPP-MicroResp. a) SEI (synthetic enzymatic index) and b) SIR (substrate induced respiration) distributions in the three soil categories (F, forest, A, agricultural and EC, extreme conditions soils)

Figure 2: Distribution of the two standardized indexes: a) H'EA and b) H'MR

a)



b)

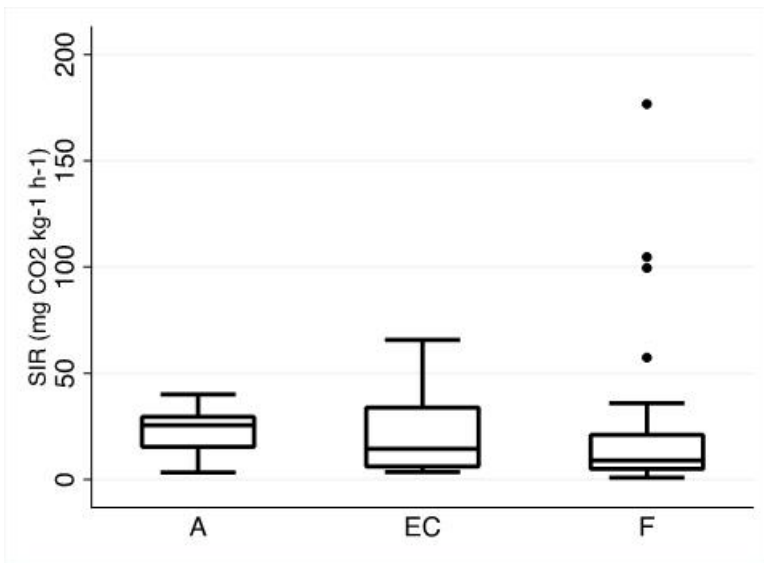
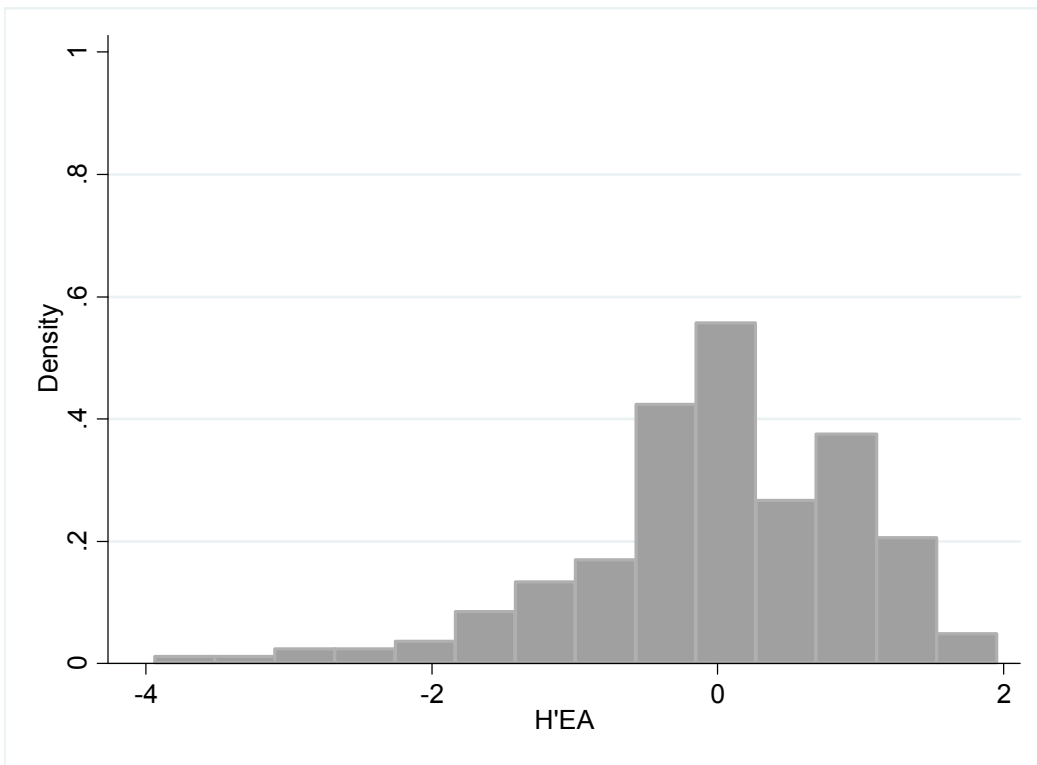


Figure 1

a)



b)

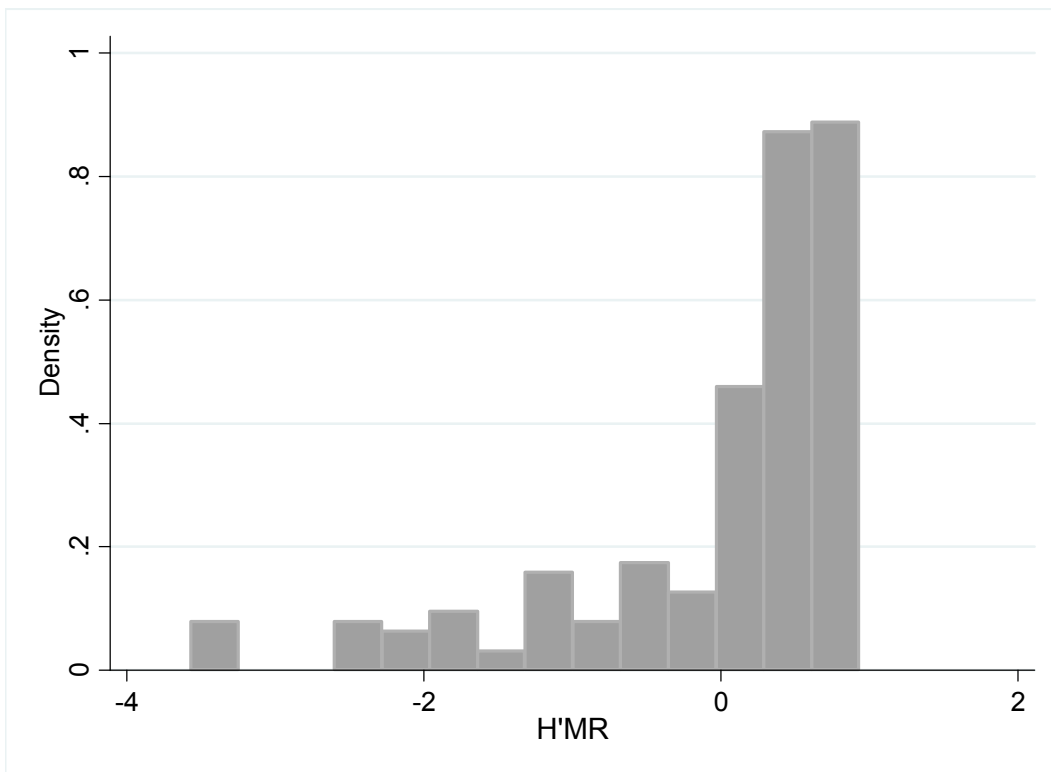


Figure 2

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1 **Table 1:** Description of all data sources. All soils are grouped into three categories: F (Forest soils), A (agricultural soils) and EC (extreme conditions soils). For  
2 each case study the following data are reported: factors of variation analyzed, soil texture, total organic carbon (TOC), pH<sub>H2O</sub>, standardized Shannon diversity  
3 index (H') measured by means of enzyme activities (H'EA) and MicroResp (H'MR), total number of samples, reference to data source. Org/conv/biodyn:  
4 organic, conventional or biodynamic management, n = number of samples. When reference was not available, a specific acknowledgement to research funds was  
5 added. Average data are reported with standard errors. n.p.=data not published.

Soil category	Factor of variation	Soil texture	TOC (%)	pH	H' EA	H' MR	n	Location	Reference or acknowledgement
Forest (F)	Management (coppiced/aged coppice)	Loam	5.9±0.5	6.4±0.1	0.61±0.01	1.11±0.01	12	Central Italy – Umbria region	Pignataro et al., (2012)
	Lithological substrate	Loam	4.2±0.7	5.8±0.2	0.78±0.02	1.07±0.04	10	Central Italy – Lazio/Umbria region	Pignataro et al., (2011)
	Afforestation ( <i>Beech</i> and <i>Douglas-fir</i> )	Sandy-clay-loam, Loamy sand	3.1±0.8	5.6±0.2	0.64±0.01	1.12±0.01	8	Central Italy – Emilia Romagna region	Papp R. (2016)
	Chronosequence ( <i>Douglas-fir</i> )	Sandy-loam, Loam	4.5±0.5	4.8±0.1	0.67±0.01	1.09±0.03	7	Central Italy – Tuscany region	Papp R. (2016)
Agricultural (A)	Management (org/conv)	Sandy-loam, Loam	1.5±0.0	7.2±0.0	0.59±0.01	1.14±0.00	30	Central Italy, Lazio region	Brunetti P. (2014)
	Management (tillage level)	Loam, Silt loam, Clay	1.9±0.2	7.2±0.3	0.55±0.02	1.08±0.01	12	North Morocco, Central Italy, Switzerland, United Kingdom, Sweden	Papp R. (2016)
	Vineyard (natural green cover/no cover)	Silty-loam	1.5±0.1	8.1±0.0	0.70±0.01	1.01±0.02	19	Northern Italy, Piemonte region	n.p.
	Management tomato crop (organic/conv)	Clay-loam	1.4±0.1	6.6±0.1	0.58±0.01	1.13±0.01	5	Central Italy, Lazio region	n.p.
	Vineyard (biodyn/conv)	Clay	1.3±0.1	6.7±0.2	0.59±0.01	0.97±0.09	4	Central Italy, Lazio region	n.p.
	Thallium contamination	Loam	6.3±0.8	6.4±0.6	0.67±0.03	1.08±0.02	8	Central Italy, Emilia Romagna region	n.p.
Extreme conditions (EC)	Arsenic contamination	Sandy-loam,Loam	2.7±0.3	5.7±0.2	0.61±0.01	1.02±0.03	14	Northern Italy, Piemonte region	Stazi et al. (2017)
	Highly calcareous, different plant cover	Sandy-loam	1.2±0.1	8.0±0.0	0.47±0.02	1.10±0.01	12	Central Italy, Lazio region	Italian PRIN 20082FC352_002
	Hydromorphous and subaqueous	Sandy	2.2±0.6	8.2±0.2	0.40±0.03	1.04±0.03	16	Central Italy, Emilia Romagna region	Papp R. et al., (2015)
	Waterlogged rice paddies and arsenic	Silty-loam, Clay-loam	1.2±0.1	7.3±0.1	0.69±0.02	0.84±0.03	20	Bangladesh	Italian PRIN 2010JBNLJ7_006
Phytoremediation (heavy metals)	Clay-loam	1.4±0.0	7.9±0.2	0.76±0.01	1.10±0.00	19	Central Italy, Tuscany region	Emili L. (2013)	

**Table 2:** Descriptive statistics of the two standardized measures. H'EA and H'MR: Shannon diversity index calculated by means of enzyme activities and MicroResp, respectively, over 196 soil samples.

Measure	Min	q0.25	q0.50	q.0.75	Max	Skewness	Kurtosis
H'EA	-3.940	-0.451	0.039	0.754	1.953	-0.851	4.279
H'MR	-3.567	-0.242	0.382	0.651	0.933	-1.746	5.506

**Table 3:** Spearman  $r_s$  values for both standardized indices H'EA and H'MR calculated for all data and within the three soil categories F, A and EC soils. ns: not significant, \*\*p<0.01

	All data	(F)	(A)	(EC)
H'EA - H'MR	-0.1183 ns	-0.0273 ns	-0.3180**	-0.0482 ns

**Table 4:** Results of Chi-squared statistics  $X^2$  and p-values obtained with Kruskal-Wallis rank test on soil functional diversity (H'EA and H'MR) among the three land use categories within restricted classes of TOC and pH. P values are reported in parentheses.

<b>TOC values</b>	<b>H'EA</b>	<b>H'MR</b>
<i>Low: TOC&lt;1.5%</i>	2.202 (0.333)	11.039 (0.004)
<i>Medium: 1.5≤TOC&lt;3%</i>	7.640 (0.022)	6.272 (0.043)
<i>High: TOC ≥3%</i>	4.431 (0.035)	7.150 (0.007)
<b>pH values</b>		
<i>pH&lt;6.5</i>	11.843 (0.003)	8.971 (0.011)
<i>6.5≤pH&lt;7.4</i>	13.867 (0.001)	30.998 (0.000)
<i>pH ≥7.4</i>	1.046 (0.306)	2.517 (0.113)

**Table 5:** Estimation results of QRMs (0.25, 0.50, 0.75 quantiles). SE= standard error, \* Significant at 5%, \*\* 1% and \*\*\* 0.1% level.

	H <sup>'</sup> EA			H <sup>'</sup> MR		
	Coef.	SE	Sign.	Coef.	SE	Sign.
<b>Quantile 0.25</b>						
TOC values ( <i>ref. Low: &lt;1.5%</i> )						
Medium: 1.5≤TOC<3	0.027	0.120		0.505	0.338	
High: TOC ≥3%	0.004	0.150		0.744	0.436	
pH ( <i>ref. pH&lt;6.5</i> )						
6.5≤pH<7.4	-0.177	0.077	**	0.295	0.183	
pH ≥7.4	-0.871	0.254	**	-0.471	0.343	
Land use category ( <i>ref. EC</i> )						
F	0.462	0.160	**	0.821	0.417	
A	0.193	0.175		0.996	0.497	*
Constant	-0.487	0.125	***	-1.276	0.426	**
<b>Quantile 0.50</b>						
TOC values ( <i>ref. Low: &lt;1.5%</i> )						
Medium: 1.5≤TOC<3	-0.197	0.183		-0.043	0.097	
High: TOC ≥3%	-0.583	0.340		-0.031	0.271	
pH ( <i>ref. pH&lt;6.5</i> )						
6.5≤pH<7.4	0.066	0.259		0.213	0.088	*
pH ≥7.4	0.417	0.342		-0.060	0.176	
Land use category ( <i>ref. EC</i> )						
F	0.415	0.305		0.418	0.230	
A	-0.377	0.180	*	0.259	0.114	*
Constant	0.260	0.307		0.171	0.120	
<b>Quantile 0.75</b>						
TOC values ( <i>ref. Low: &lt;1.5%</i> )						
Medium: 1.5≤TOC<3	-0.001	0.142		0.008	0.072	
High: TOC ≥3%	-0.549	0.256	*	-0.070	0.136	
pH ( <i>ref. pH&lt;6.5</i> )						
6.5≤pH<7.4	-0.008	0.233		0.079	0.073	
pH ≥7.4	0.527	0.235	*	-0.096	0.116	
Land use category ( <i>ref. EC</i> )						
F	0.422	0.192	**	0.239	0.111	*
A	-0.703	0.136	***	0.160	0.080	*
Constant	0.804	0.246	**	3.212	0.414	**

Notes: \* p<0.05; \*\* p<0.01; \*\*\*p<0.001

