

1 **Hydrogeological and geochemical characterization of groundwater in the F’Kirina**
2 **plain (Eastern Algeria)**

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18 **Abstract**

19 The F’Kirina plain in eastern Algeria is an endorheic basin suffering water scarcity due to a
20 combination of natural and man-made causes. Its hydrogeological system is complex as
21 made by interconnected aquifers represented by Mesozoic, Cenozoic, and Quaternary
22 lithological units. The combination of drought indicators and water level data shows that a
23 groundwater drought affected the plain during the last 15 years, which reflects on current
24 water quality. The reported geochemical analyses, including major ions and trace
25 elements, indicate that the groundwater resource is suffering from salinization, mainly due
26 to evaporation and leaching of soil salts, a process that is coupled with simultaneous
27 cation-exchange effects. In this framework, we observe a geochemical evolution from the
28 fresh Ca-HCO₃ facies, typical of springs bordering the plain, towards more saline
29 groundwater characterized by chloride/sulphate-rich facies in the middle of the plain
30 approaching the sebka. However, geochemical diagrams indicate that in few wells
31 salinization is also influenced by upraising of deep groundwater. The water isotopic
32 composition of the F’Kirina plain samples suggests that they diversely record both
33 recharge and evaporation components. Moreover, the most ¹⁸O and D depleted
34 compositions among the investigated ground-waters suggest recharge contributions by
35 comparatively higher elevation or the involvement of old (fossil) water components.

36 **Keywords**

37 Groundwater salinization; Geochemical tracers; Stable isotopes; Drought; Water scarcity;
38 Endorheic basin

39 **1. Introduction**

40 Water resources in the arid and semi-arid regions of the world are characterized by
41 scarcity and high spatial and temporal variability. In these regions, with population growth,
42 expansion of irrigated areas, and climate change, there is an increasing groundwater
43 demand, which is the major source of water for drinking, irrigation, and industrial uses
44 (UNESCO 2020).

45 North Africa, including Algeria, is considered a “climate change hotspot” (Diffenbaugh and
46 Giorgi 2012) with a high year-to-year variability of rainfall amounts, drought periods, and
47 heat waves (Cook et al. 2016; Lelieveld et al. 2016, Lionello 2012; Mariotti et al. 2015).
48 Many projections and models estimate for the investigated area (and surroundings) a total
49 annual precipitation decrease of 15-30% by the end of the twenty-first century
50 (Christensen et al. 2007; Hadour et al. (2020) and that droughts will become more
51 frequent and severe than those already observed during the 20th century (Reiser and
52 Kutiel, 2011; Sousa et al., 2011; Hoerling et al., 2012; IPCC, 2013; Schilling et al., 2012;
53 Zeroul et al., 2019).

54 Consequently, serious questions arise about the future agricultural practices, food safety,
55 and local population displacement in Algeria and surrounding areas in the Middle East and
56 North Africa (MENA) region (Alboghday and El-Hendawy, 2016).

57 On this framework, other researchers focusing attention on water scarcity in the MENA
58 region emphasized that water over-exploitation is aggravating the climatic effects
59 (Haddadin, 2001; Leduc et al., 2017). Indeed, in semi-arid and arid regions of North Africa,
60 long-term and large-scale exploitation of groundwater have significantly contributed to the
61 deterioration of aquifers, with groundwater level decline (Bouchaou et al., 2011) and
62 concurrent salinization, leading to the deterioration of the groundwater-dependent
63 ecosystems.

64 Salinization often originates from a combination of natural and human factors, including
65 seawater intrusion, dissolution of evaporite minerals (Bouchaou et al 2008; Re et al.,
66 2014), mixing with deep saline waters/brines and geothermal fluids, cycling wetting and
67 drying, and agricultural practices (irrigation and use of fertilizers). Seawater intrusion
68 (accompanied by water-rock interaction) is often the main mechanism controlling
69 groundwater salinization in North African coastal aquifers, as in Jerba Island (Tunisia,
70 Souid et al., 2020) or Korba coastal plain (Tunisia, Zghibi et al., 2013); however, this is not
71 the case of the F’Kirina inland aquifer. It is important to note that salinization affects also
72 North Africa’s inland aquifers far from coastal zones. In this framework, salinization is
73 mainly due to the dissolution of evaporite salts, as observed in several North African inland
74 aquifers and endorheic basins (Farid et al., 2015; Mejri et al., 2018; Bouragba et al., 2011;
75 Zereg et al 2018; Aouidane and Belhamra 2017).

76 Many inland basins in North Africa are characterized by the presence of chotts and
77 sebkhas that form as salt pans, playa systems, and saline ephemeral lakes on low-laying
78 and internally drained areas because of the local topography and the prevailing desert
79 conditions. Information on saline waters contained in the chotts located in eastern Algeria
80 comes from studies by Merzouk et al (2020) on Sahara lithium resources. The authors
81 described that the surface of chotts is composed of halite and gypsum crusts, clastic
82 muds, and very fine aeolian powder. The chott water levels mainly depend on the complex
83 interplay between groundwater and surface water recharge, and subsequent evaporation.
84 Coherently, the chott waters are often brines, where the effect of evaporation seems
85 stronger than that of dilution by recharge from aquifers and precipitation; the main source
86 of major ions is the dissolution of salts such as halite, epsomite, and gypsum. Chotts have
87 a role in conditioning the hydrogeology of concerned basins and the quality of
88 groundwater; many authors emphasize Total Dissolved Solid (TDS) gradients towards
89 chotts and sebkas, likely caused by salt dissolution (Belkhiri et al. 2012; Ghodbane et al.,
90 2015; Aissa et al., 2017), and/or by groundwater evaporation from salt pans as described
91 by Schulz et al., 2015.

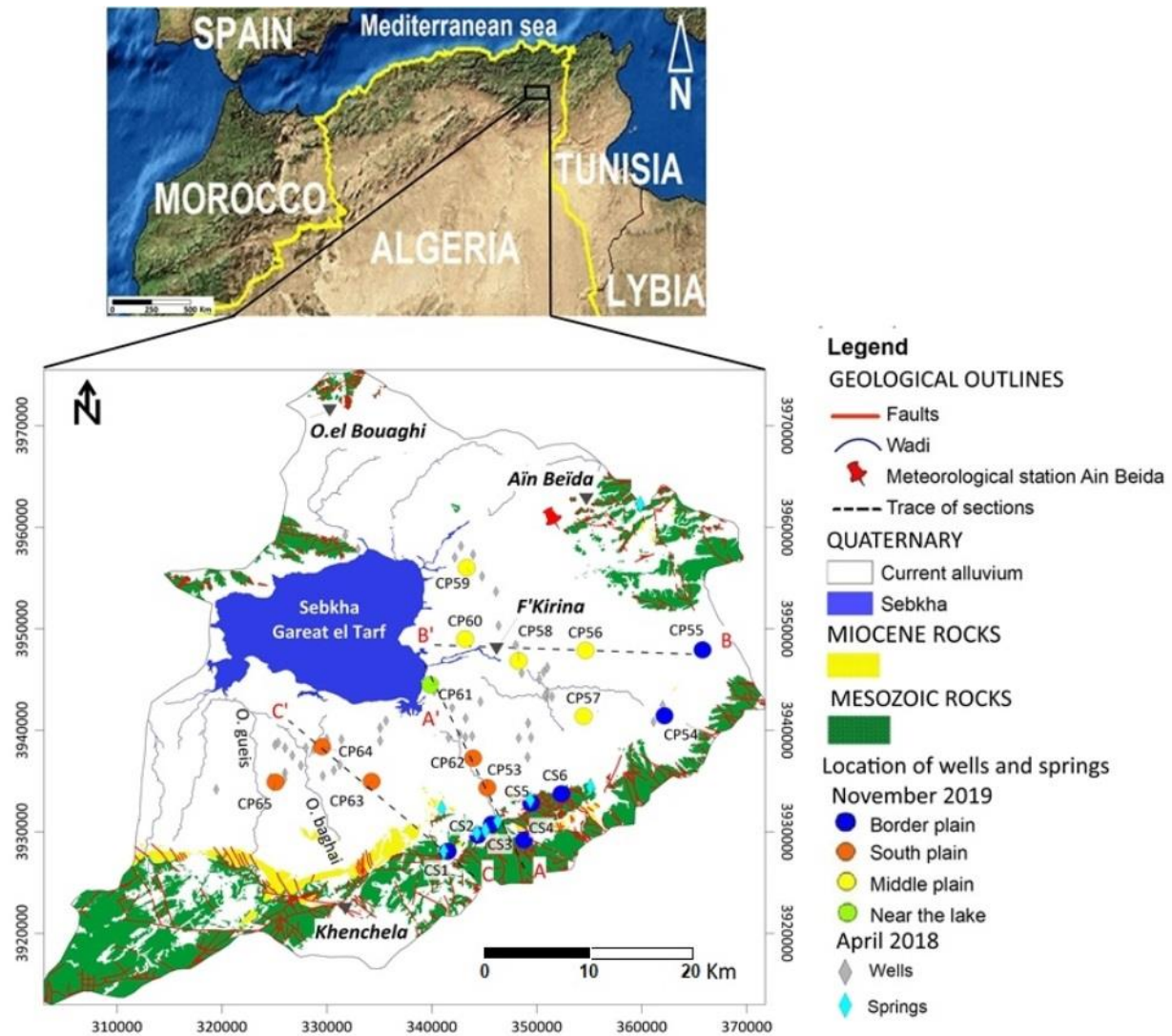
92 Noteworthy, the available studies on groundwater salinization rely on isotope hydrology
93 and/or geochemical methods to recognize groundwater origin, sources of salinization, and
94 associated water-rock interaction processes.

95 In this light, our research concerning the groundwater resources of the F'Kirina plain
96 (north-eastern Algeria; Fig. 1) addresses the problem of groundwater salinization based on
97 new hydrogeological, geochemical, and isotope data on groundwater obtained during two
98 surveys, in April 2018 and November 2019.
99 These data are interpreted considering the hydrogeological structure of the aquifer and
100 climatic conditions and give clues on the origin of groundwater recharge, flow paths, and
101 water-rock interactions. Results are finally discussed in terms of source(s) and processes
102 of groundwater salinization.

103 **2. Description of the study area**

104 105 *2.1. Geographical and geo-hydrogeological setting of the study area*

106
107 The F'Kirina plain, extended over an area of 650 km², is located in the North-East part of
108 Algeria (North Africa) at the midpoint between the Mediterranean Sea and the Sahara (Fig.
109 1). It is part of the Constantine high plain and belongs to the Gareat El Tarf basin that is an
110 extensive endorheic depression with a mean elevation of 960 m above mean sea level
111 (AMSL), surrounded by mountains whose summits exceed the 2000 m AMSL.
112 The southern and northern borders of the basin are marked by relatively steep slopes as a
113 consequence of the tectonic activity, which affected the region. The existing streams (Nini
114 wadi, Ouelmene wadi, and Isfer wadi) flow toward depressions such as the Chott Gareat
115 El Tarf, which is partially dry and salt-encrusted. This Chott is a protected wetland
116 classified as Ramsar site and merges with other minor chotts located westward during
117 major floods.



118 Fig. 1. Geological schematic map of F'Kirina plain, including the location of April 2018 and November 2019
 119 sampling points; the latter are distinguished by geographical zonation (Coordinate Reference System: Datum
 120 World Geodetic System 1984, WGS 84, UTM ZONE 32N)
 121

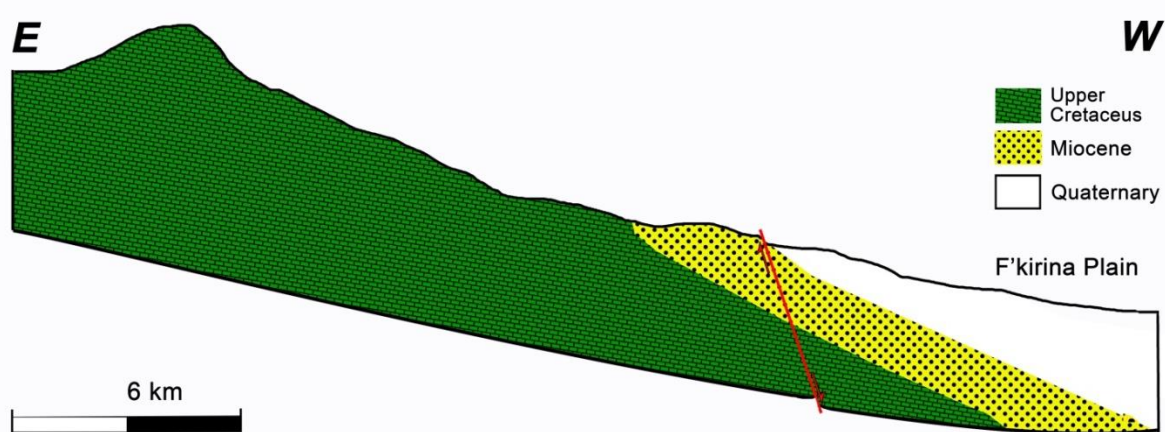
122 The geology of the Gareat El Tarf basin (which includes the F'Kirina plain) has been
 123 described by Voute (1967), Vila (1977), Guiraud (1977), and Benazzouz (1986). The
 124 stratigraphic series ranges from the Triassic to the Quaternary (Wilidi 1983). Formations
 125 of Mesozoic and Cenozoic age locate to the extreme East, North-East and South of
 126 F'Kirina plain (Fig. 1). The area characterized by Mesozoic rocks includes formations from
 127 Triassic to Upper Cretaceous. The latter form the backbone of the region: they are roughly
 128 oriented SW-NE and are largely formed by limestone and marl. Cenozoic (mainly Mio-
 129 Pliocene) formations characterise the middle and the south plain, where Quaternary
 130 formations also outcrop. The Sebkhia of Gareat El Tarf is marked by Triassic outcrops,
 131 Miocene conglomerates, and by Quaternary deposits. The silty and clayey soil,
 132 impregnated by salts, is of low permeability.

133 The structures and landforms are the results of the two main Tertiary orogenic phases that
 134 are the Atlasic phase of Middle–Late Eocene age and subsequent Alpine Miocene phase
 135 (Laffitte 1939; Durosoy 1956). The Atlasic phase gave rise to large SW-NE and ENE-
 136 WSW folds, which were reactivated during the post-Miocene phases producing NW-SE
 137 and NE-SW faults (Guellal and Vila, 1973). The Miocene sediments lay on erosion
 138 surfaces thus forming a series of sometimes very deep depressions (Laffitte, 1939; Marmi,
 139 1995). During Late Quaternary, E-W to NE–SW striking folds and reverse faults affected

140 the young Quaternary deposits (Harbi et al, 1999). In particular, a syncline with younger
141 Cenozoic rocks in inner parts and older Mesozoic rocks at the periphery is observed in the
142 studied area.

143 From the hydrological point of view, the Lower Maastrichtian marly substratum outcropping
144 at the border of the plain constitutes an impermeable base that is overlaid by permeable
145 lithologies. The former includes the Maastrichtian fractured and karstified limestone
146 outcropping to the East (whose thickness should not exceed 250 m) and the coarse
147 Tortonian (Miocene) sandstones; the limestone of the Upper Maastrichtian constitutes the
148 deep aquifer for the F'Kirina plain. The shallower hydrogeological structure corresponds to
149 alluvial heterogeneous sediments (Mio-Plio-Quaternary) of variable permeability having in
150 the whole a thickness of more than 300 m (Younsi, 2009). The above hydrogeological
151 elements are synthesised in the schematic E-W geological section of Fig. 2. The main
152 aquifer within this structure, which constitutes the most important source of groundwater
153 for the area, is the Plio-Quaternary aquifer, with a large horizontal extension and a
154 thickness from 10 to 100 meters, characterized by a succession of heterogeneous fluvial-
155 lacustrine sediments composed of alternating gravels and clays, calcareous gravels, and
156 pebbles.

157 The heterogeneity of the Plio-Quaternary lithology gives origin to a multi-layered aquifer
158 system. Technical sheets of drillings (courtesy of Dept. of Water Resources, Wilaya of
159 Khenchela) highlight a succession of saturated levels; local disconnection of aquifer levels
160 is due to lenses of clays and marls, often with poor lateral continuity. However, the vertical
161 hydraulic connection among saturated layers may naturally occur.



162 Fig. 2. Schematic E-W geological section

163 The Plio-Quaternary aquifer, beside a direct and local recharge from precipitation, receives
164 a contribution from the Maastrichtian (Cretaceous) fractured and karstified limestone
165 aquifers outcropping to the East and possibly from the Aures massifs to the South.

166 3. Material and methods

167 3.1 Precipitation data and groundwater level series

168 The study considers monthly average precipitation and temperature series measured in
169 the period 1987-2020 at the rain gauge station of Ain Beida, located to the northern border
170 of F'Kirina plain (Fig.1) (courtesy of National Meteorological Office of Ain Beida, Oum el-
171 Bouaghi, Algeria). Data has been elaborated to calculate the meteorological drought
172 indicators SPI (Standardized Precipitation Index, McKee et al. 1993) and SPEI (Standard
173
174

175 Precipitation-Evapotranspiration Index; McKee et al., 1993; Guttman, 1999; Lloyd-Hughes
176 and Saunders, 2002; Vicente-Serrano et al., 2010; Beguería et al., 2014) at 12-month
177 scale, which reflects medium-term precipitation patterns, thus being of hydrological
178 significance (Potop et al 2012, 2014).

179 Time trends of the SPI and SPEI indicators have been compared to those of historical
180 water level data (period 2003-2018) collected about every six months during low and high
181 water periods by the Agence Nationale des Ressources Hydriques (ANRH) of Constantine
182 (Algeria) on a monitoring network of 21 wells located in the F'Kirina plain. Measures were
183 made with conventional hand-held devices.

184

185 *3.2 Field surveys and sampling point technical data*

186

187 The first survey concerning water level measure was carried out in April 2018 on a net of
188 73 wells (location in Fig. 4). On the same date, water sampling for preliminary (major
189 element) analysis was carried out at 55 out of 73 wells and at 9 springs (Fig. 1). The
190 investigation was repeated in November 2019, selecting 13 wells and 6 springs (location in
191 Fig. 1) for a more complete geochemical investigation, which is presented in this paper.
192 Most of the wells involved be the November 2019 sampling survey are located along the
193 sections AA', BB' and CC' (Fig. 1).

194 Geographical and technical data of sampling points of this survey are in Tables 1S and 2S
195 (Supplementary material). Water samples are grouped according to their geographical
196 position as belonging to *i*) Border Plain (2 wells and 6 springs located at the highest
197 elevation between 912 and 1282 m AMSL), *ii*) South plain (5 wells), *iii*) Middle plain (5
198 wells), and *iv*) Near the lake (1 well). The use of ground-waters from wells is both for
199 domestic and irrigation purposes, while the main use of spring waters is domestic.

200 The considered wells are mainly drilled, with a mean depth of 107 m below ground (b.g.).
201 Wells use shallow tube well technology, combined with small engine-driven pumps;
202 although there is no information about the depth reached by tubes, the length of wells
203 (Table 2S, Supplementary Material) indicates that they roughly cross the whole thickness
204 of the Plio-Quaternary lithologies.

205 Most well technical sheets indicate a succession of saturated levels, suggesting that
206 groundwater samples drawn by pumping may collect groundwater from more than one
207 aquifer level, thus giving potential "mixed" information concerning their quality.

208

209 *3.3 Water sampling and geochemical analyses*

210

211 Groundwater samples in both sampling surveys were taken after well purging for at least
212 20 minutes; then, pH, Temperature, and Electrical Conductivity (EC) were measured in
213 situ on each sample, using Portable Analyser Kit (Hanna HI 9812-5). Samples were stored
214 at 4°C. Alkalinity was measured within a day at the Laboratory of Sedimentary
215 environment and mineral and water resources in eastern Algeria of Tebessa University
216 (Algeria). Samples collected in April 2018 were analysed at the same laboratory for a
217 preliminary characterization concerning the major ions composition. Ca²⁺, Mg²⁺ were
218 analysed by atomic absorption spectroscopy using an Analyst 400 (Perkin Elmer), and Na⁺
219 and K⁺ with flame photometer model 410 (Sherwood); the volumetric method was used for
220 bicarbonates and chlorides, while the spectrophotometer DR2000 (Hach) was used for
221 sulphates. The average error balance for the 2018 samples is 1.3%.

222 Samples collected in November 2019 were analysed in Italian laboratories for a detailed
223 investigation including minor and trace elements as well as oxygen and hydrogen
224 isotopes. Anions (F⁻, Cl⁻, Br⁻, NO₃⁻, SO₄²⁻) were analysed at the Department of Physics and
225 Earth Sciences of the University of Ferrara (Italy) by ion chromatography using a Dionex

226 ICS-1000. Cations (Ca^{2+} , Mg^{2+} , Na^+ , K^+) were analysed by ion chromatography using
227 Metrohm IC-900 at the Department of Land, Environment, Agriculture, and Forestry of the
228 University of Padova (Italy). The dissolved silica concentration (SiO_2 in mg/L) was
229 determined using the Silica Portable Photometer HI-96705 (Hanna Instrument) that works
230 in the range 0-2 mg/L. The average error balance is -2.3%.

231 Trace elements were analysed through an Agilent 7800 ICP-MS at the Department of
232 Earth Sciences of the University of Florence (Italy) using Rh as internal standard, the
233 Agilent Technologies tune solution containing 1 $\mu\text{g/L}$ of Ce, Co, Li, Mg, Y in 2% HNO_3
234 matrix, and the Merck VI multi-elemental standard solution for calibration. Samples have
235 been analysed at different dilutions to match the calibration range for most elements.
236 Accuracy and precision, calculated on the basis of repeated analyses of samples and
237 standards, were better than 10 % for all the considered parameters.

238 Hydrogen and oxygen isotopic composition was determined at the Physics and Earth
239 Sciences Department of the University of Ferrara, using the CRDS Los Gatos LWIA 24-d
240 isotopic analyser (Los Gatos Research). The isotopic ratios of $\text{D}/^1\text{H}$ and $^{18}\text{O}/^{16}\text{O}$ were
241 expressed as δ notation ($\delta = (R_{\text{sample}} / R_{\text{standard}} - 1) \times 1000$) to the Vienna Standard Mean
242 Ocean Water (V-SMOW) international standard. Three bracketing standards were
243 systematically run during the analytical sessions. These standards, obtained from the Los
244 Gatos Research Company, were calibrated according to the international IAEA
245 (International, Atomic Energy Agency) protocol. Analytical precision and accuracy, based
246 on replicate analyses of standards, were better than 0.3% and 1.0% for $\delta^{18}\text{O}$ and $\delta^2\text{H}$,
247 respectively (Marchina et al., 2015; Natali et al., 2016). As indicated in Marchina et al.
248 (2020), the average standard deviation of 366 samples is 0.7‰ for δD and 0.14‰ for $\delta^{18}\text{O}$.
249 Regression lines calculated in the next sections are obtained by the Ordinary Least
250 Squares (OLS) method, and no significant differences are observed adopting other
251 regression methods (Marchina et al., 2020).

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253 **4. Results and discussion**

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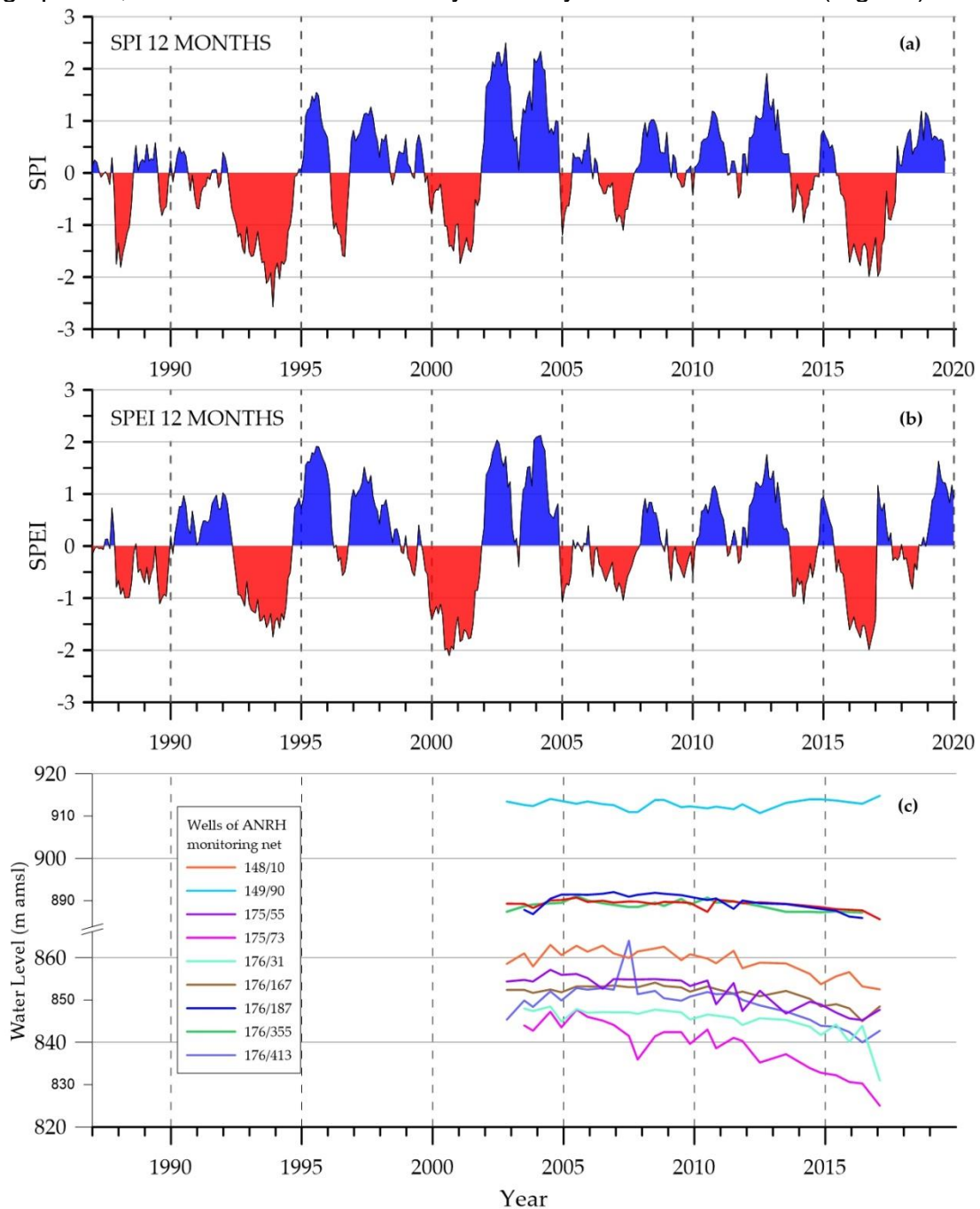
255 *4.1 Climatic and hydrogeological setting*

256

257 The results of the SPI and SPEI calculation at 12-month scale are shown in Fig. 3a and 3b
258 respectively. SPI and SPEI give a different evaluation of drought intensity, however
259 highlighting the same periods as to their occurrence. For the following considerations, we
260 will consider the SPEI index, considered more reliable than SPI in the light of the inclusion
261 of temperature in the calculation. Table 1 shows, concerning the different SPEI categories
262 of drought/wet severity, the number of months in the period 1987-2020, and the
263 percentage of months in the two periods 1987-2003 and 2004-2020.

264 Both Fig. 3b and Table 1 indicate that in the period 2004-2020 compared to the 1987-2003
265 period, there has been a decrease in the percentage of moderate, severe, and extreme
266 drought months as well as of extremely, severely and moderate wet months, with an
267 increase of the percentage of “normal months”. In the whole period of analysis, 50.5% of
268 months shows a $\text{SPEI} < 0$; the comparison between the two sub-periods indicates a
269 decrease of the total number of months with $\text{SPEI} < 0$ during 2004-2020. Thus, the F’Kirina
270 plain has suffered from droughts much more in the past than in the more recent period.
271 The SPEI values indicate that the years of sampling (2018 and 2019) are “normal”;
272 however, they are preceded by a period of severe meteorological droughts (spring 2016-
273 winter 2017). The only data useful to understand the reaction of F’Kirina plain groundwater
274 to the succession of drought periods are those collected for the period 2003-2018 by the
275 ANRH.

276 Fig. 3c shows the time series of 9 (location in Fig. 4) out of the 21 wells of the net because
 277 they are the only ones for which measurements are regular on a half-year basis. However,
 278 selected time-series cover the level range characterizing the eastern part of the study
 279 area. The considered time-trends show that, after the severe drought of the period 2000-
 280 2002, WLs were steadily increasing till the middle of 2005 under the influence of the wet
 281 period 2003-2005. Then, from this period onwards, WLs, under a sequence of normal
 282 periods, gradually lowered. They reached their lowest levels in January 2018 (last
 283 measure) after the severe drought period of 2015-2017: compared to the maximum levels
 284 measured in mid-2005, the WLs showed a decrease between 5 and 23 m. Other records
 285 (not included in Fig. 3c because discontinuous) showed that at the end of this drought
 286 period some wells were completely dry. The last drought period is followed by a low
 287 recharge period, which became moderately wet only at the end of 2020 (Fig. 3b).



288
 289 Fig. 3. SPI (a) and SPEI (b) calculated on a 12-month scale; (c) historical water level time-trends (ANRH,
 290 location of the wells in Fig. 4)

Table 1: Distribution (%) of different types of droughts according to SPEI

Period	% of months with SPEI < 0 on total months in the period	
1987-2020	50.2	
1987-2003	52.2	
2004-2020	48.7	

Categories of drought/wet severity	SPEI value	Number of months 1987-2020	Period 1987-2003	Period 2004-2020
			% of months/total months	% of months/total months
Extremely wet	$SPEI \geq 2$	7	0.5	2.7
Severely wet	$1.5 \leq SPEI < 2$	23	6	5.9
Moderate wet	$1 \leq SPEI < 1.5$	53	11	8.9
Normal	$-1 < SPEI < 1$	410	61.7	70.1
Moderate dry	$-1.5 < SPEI \leq -1$	41	13.9	7.5
Severe drought	$-2 < SPEI \leq -1.5$	18	5.9	4.9
Extreme drought	$SPEI \leq -2$	4	1.5	0.5

292 The available data allow evaluating, the “groundwater drought” from 2005 till 2018,
 293 following the first severe drought period of 2000-2002. Groundwater droughts (with effects
 294 on groundwater quality and quantity) are consequent to meteorological droughts but are
 295 generally delayed compared to the onset of superficial effects of water depletion (Robins
 296 et al., 1997). Lags mainly depend on the severity and duration of meteorological droughts,
 297 their time sequence, physical characteristics of aquifers, and amount groundwater demand
 298 during rainfall breakdown. In the light of the above considerations and despite the low
 299 frequency of water level measures, we can conclude that both the investigated years 2018
 300 and 2019 were under the effect of a long groundwater drought, with an influence on
 301 groundwater geochemical features.

302 The number of water level measures of April 2018 and the distribution of wells in the space
 303 was adequate to allow the elaboration of WL contour lines by Ordinary Kriging (Fig. 4).
 304 Water levels varied from more than 900 to less than 830 m AMSL, outlining a direction of
 305 groundwater flow from the recharge areas located in the eastern and southern mountains
 306 towards the Sebka of Gareat El Tarf. Water levels measured in October 2019 varied in the
 307 same range as those of April 2018, but either the number or the spatial distribution of
 308 prospected wells do not allow for meaningful geostatistical processing.

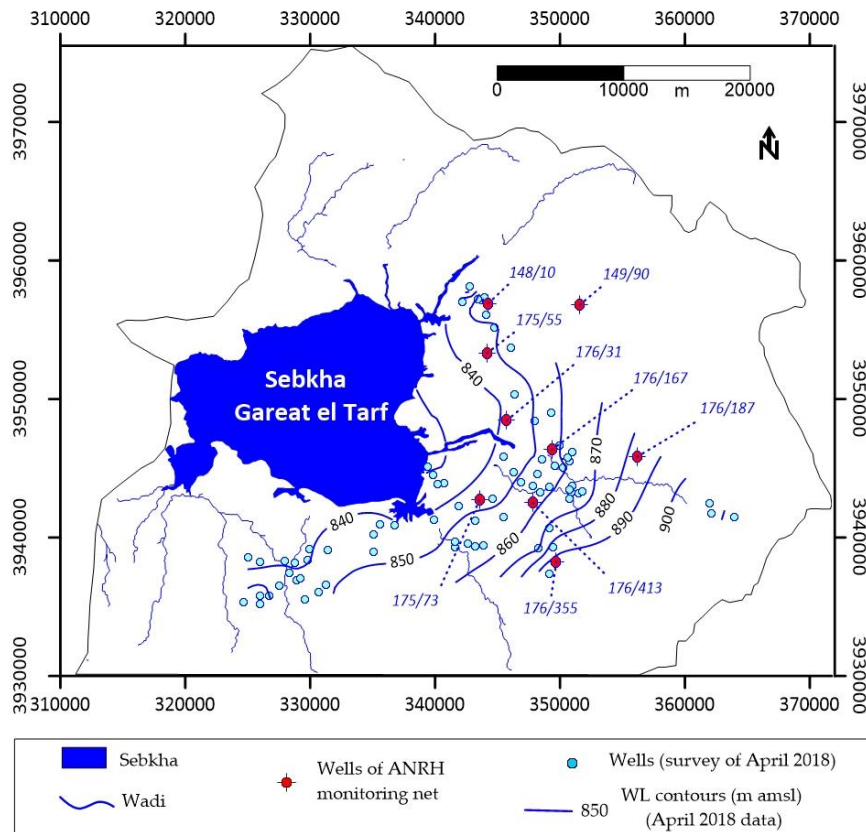


Fig. 4. Water level (WL) contours line of April 2018; location of wells of the ANRH monitoring net

309 **4.2 Geochemical characterization of the studied waters**

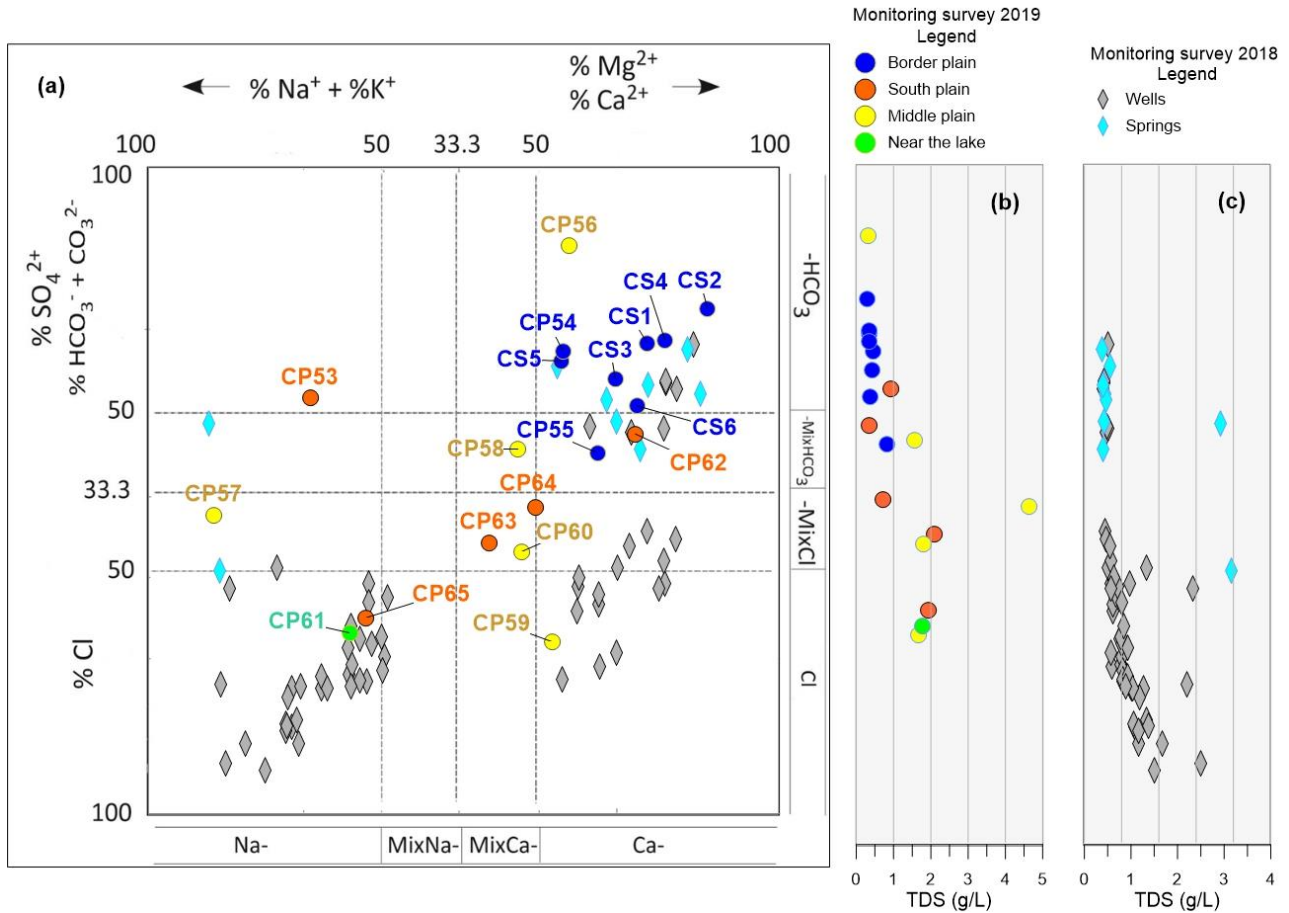
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311 The data relating to groundwater and spring samples collected in November 2019 are
 312 discussed given their geographical zonation (Fig. 1); they are compared to data
 313 concerning the 2018 samples only for major elements. The whole data-set is commented
 314 in the light of the previously described hydrogeological and conceptual information.
 315 Analytical results of the 2019 survey, including the physicochemical measurements carried
 316 out in the field and the geochemical analyses, are reported in Table 3S (Supplementary
 317 material).

318 Preliminary statistical treatment of the 2019 analytical results has been obtained through a
 319 cluster analysis (dendrogram in Fig. 1S, Supplementary Material); the hierarchical
 320 relationship between samples highlights a link between the spatial distribution and the
 321 geochemical features, with some outliers (e.g. CP57, and in the minor extent CP53) that
 322 are quite different from the main group of samples. Results also show that some tracers
 323 such as chloride, bromide, sulphate, and sodium appear well correlated ($R^2 > 0.7$, see the
 324 correlation plot in Fig. 2S, Supplementary Material) showing that their concentration is
 325 regulated by common dissolution and/or water-rock interaction processes.

326 The content in Total Dissolved Solids (TDS) of the 2018 groundwater samples varies
 327 between 397 and 2,491 mg/L, with an average of 933 mg/L, while spring waters show a
 328 TDS varying between 375 and 3,146 mg/L, with an average of 1,016 mg/L. As concerns
 329 the groundwater samples of the 2019 survey, those of the plain show a TDS between 337
 330 and 4,701 mg/L, with an average value of 1,473 mg/L, while the TDS of the spring waters
 331 collected at the border plain varies between 303 and 466 mg/L, averaging 378 mg/L. TDS
 332 values higher than 1,500 mg/L relate to CP65 and CP63 samples in the South plain, to

333 those collected in the Middle plain (CP58, CP59, CP60, and CP57), and Near the lake
 334 (CP61).
 335 Water samples of both sampling survey have been classified (Fig. 5) using the
 336 Hydrochemical Facies Evolution diagram (HFE-D, Giménez-Forcada, 2010; Giménez-
 337 Forcada and Sánchez San Román, 2015). In this diagram, hydrochemical facies (HF)
 338 are determined as a function of calcium and sodium (cations) percentages, and bicarbonate
 339 and chloride (anions) percentage to the sum of cations and anions, respectively.
 340



341
 342
 343 Fig. 5. a) HFE diagram of the studied waters; b) and c) TDS variation of the 2019 samples and 2018 ground-
 344 waters and spring waters respectively, corresponding to HF evolution. The 2019 samples are distinguished
 345 by their code and marked according to geographical zonation (Fig.1).
 346

347 The diagram includes four main “heterotopic” facies (Na-HCO₃, Na-Cl, Ca-HCO₃, and Ca-
 348 Cl), which are defined by cation and anion percentages higher than 50%. When a facies
 349 includes the Mix term coupled to a cation or an anion or both ions (hybrid facies), this
 350 indicates that the percentage of the cation or anion is less than 50%, but, at the same
 351 time, it is the highest percentage compared to the other considered cations and anions.
 352 The abscissae, which separately represent the percentages of Na⁺ and Ca²⁺ in meq/L,
 353 reproduce the base-exchange reactions: thus, the HFE-D is particularly effective in the
 354 description of heterotopic (as Na-HCO₃ and Ca-Cl) and hybrid HFs often occurring when
 355 mixing of distinct groundwater end-members triggers cation-exchange processes. The
 356 percentages of anions are in the ordinates, where the percentage of chloride is connected
 357 to groundwater salinization due to seawater or chloride-salts and the percentage of
 358 bicarbonate or sulphate (depending on the dominant anion in freshwater) characterizes
 359 recharge waters. Fig. 5 integrates a plot showing for each HF the corresponding TDS
 360 (Giménez-Forcada, 2018). The diagram allows outlining water evolution within the aquifer.

361 The HF Ca-HCO₃ marks the 2019 springs, the samples CP54 and CP56, and some
362 springs and wells of the 2018 survey. These pristine waters evolve to CaMixHCO₃ type
363 (CP62, CP55, and other springs and wells of the 2018 survey) with TDS increase. Further
364 increase in TDS leads to hybrid HFs such as MixCaMixCl (CP60, CP63, CP64) and
365 MixCaMixSO₄ (CP58); then hydrochemical facies as NaCl appear in the Middle plain
366 (CP65) and Near the lake (CP61), maybe indicating dissolution of chloride (and sulphate)
367 salts; this HF also marks most of the samples of the 2018 survey. The remaining 2019
368 samples follow different evolution paths: the CP53 and CP57 samples show the Na-HCO₃
369 and NaMixCl facies, respectively, while the sample CP59 shows the Ca-Cl facies (as many
370 of 2018 samples). The genesis of these last HF types of groundwater is not straightforward
371 and can be related to various factors. The HF of CP53, which is quite fresh (0.9 g/L), can
372 be the proxy of a deep fresh end-member circulating in rocks including Na-rich ion-
373 exchangers (clay), where a slow equilibration of freshwater with the exchange complex
374 causes a series of reactions that favour the sequential release of Mg²⁺ and Na⁺ to the
375 solution in exchange for Ca²⁺. The concurrent increase in a closed environment of total
376 alkalinity and Na⁺ could be related to the availability of calcite, which dissolution is triggered
377 by the under-saturation induced by the loss of dissolved Ca²⁺ subtracted by cation
378 exchange. This process implies that Na⁺ amounts in the exchangers are much higher than
379 those of Ca²⁺ carried by freshwater, resulting in a very high ratio of adsorbed ions/ions in
380 solution: the succession of the different hydrochemical facies during this process of
381 softening is, therefore, a rather slow process (Lambrakis and Kallergis, 2001), since it
382 requires the passage of a volume of freshwater contributing Ca²⁺ in relatively low
383 concentration many times greater than the volume of the pores to move all the Na⁺ from
384 the exchange sites.

385 The CaCl facies of CP59 likely originates from the inverse exchange Na/Ca: such type of
386 facies is normally observed during active seawater intrusion on coastal aquifers where clay
387 sediments in equilibrium with freshwater are flushed by saline waters (Appelo and
388 Willemssen 1987; Beekman 1991). Same HFs can be originated in continental aquifers in
389 connection to groundwater salinization by the dissolution of salts.

390 The CP57 (with a TDS of 4.7 g/L) shows a NaMixCl HF; however, it has a high content of
391 bicarbonates (1500 mg/L) with a percentage barely smaller (24.6%) than that of chlorides
392 (25.6%). CP57 positions along an evolution path in between two springs (named Elkenif
393 and Gaarir) sampled during the 2018 survey: these springs are marked by NaCl, and
394 NaMixSO₄ HFs, and a TDS of 3.1 and 2.9 g/L, respectively. Both springs are thermal
395 springs used from the Roman time as Spa. They show high temperatures at the outlet (46
396 and 38°C). This analogy shows that salinization in the area is not linked only to
397 evaporation and leaching of Quaternary salty deposits at shallow levels but, at least in few
398 cases, can be inherited also by deeper contributions outpouring from the basal levels of
399 the Mesozoic rocks.

400 Thus, F'Kirina groundwater may be subject to different processes, according to the aquifer
401 lithology, as ion exchange and dissolution of salts. For comparison, we report variations in
402 salt content and facies shown by other researchers who studied the F'Kirina plain in the
403 last years. Dib et al. (2017) analysed 47 samples (unknown sampling date), classified in
404 F'Kirina and Ain Zitoune groups, which roughly correspond to the Middle and South Plain
405 groups of our study. The EC of ground-waters of F'Kirina group is between 550 and 6060
406 μS/cm, while the EC of samples of the Ain Zitoune group varies in the range 600-15,800
407 μS/cm. The dominant HFs are CaMgSO₄Cl and CaHCO₃. Dib et al. (2017) attributed the
408 significant correlation ($r^2 > 0.8$) between sodium, chlorides, and sulphates to widespread
409 evaporation processes. Salima and Belgacem (2017) analysed 45 samples (May 2015) in
410 an area corresponding to the Middle Plain. The EC ranges from 220 to 6,700 μS/cm, while
411 HFs vary from CaHCO₃ to CaSO₄ in the Plio-Quaternary sediments; CaCl facies are found

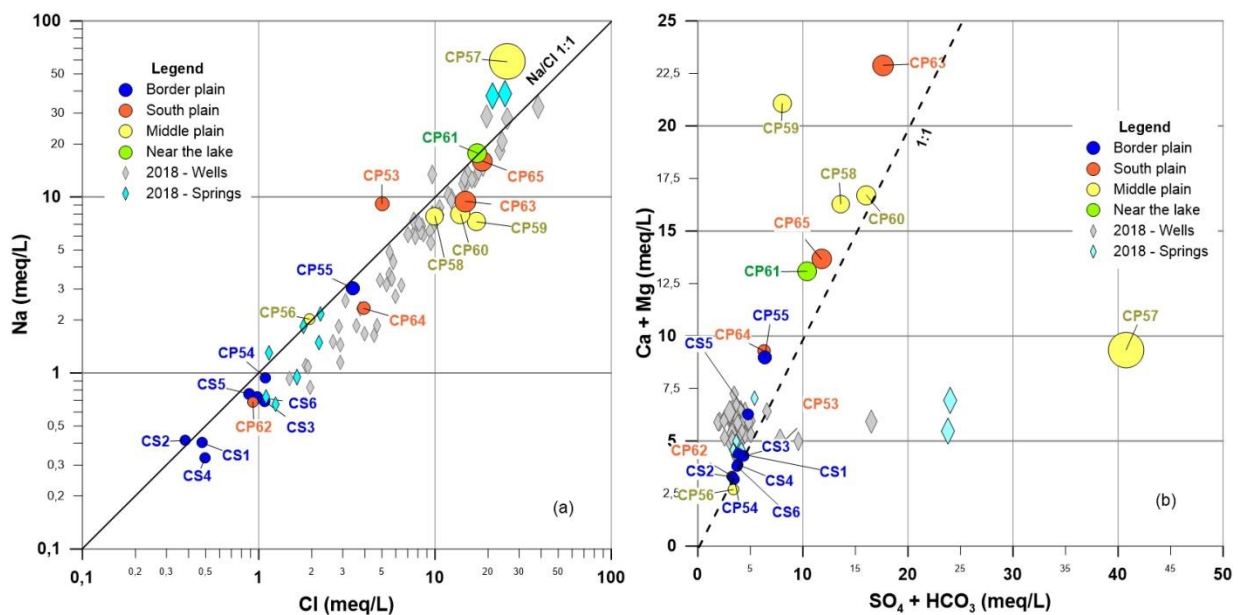
412 near the Sebkhha. In November 2016 Kahal et al. (2018) analysed 21 groundwater samples
 413 in an area corresponding to the Middle Plain; they found predominant CaSO_4 , CaHCO_3 ,
 414 and CaCl facies; there are no data on EC and TDS. These findings suggest that water
 415 quality in the F'Kirina plain is highly variable, depending on the selected well net, and
 416 survey season and year.

417 Figs 6a and 6b show the relationships between Na and Cl and (Ca+Mg) vs (HCO_3 + SO_4)
 418 respectively. The Na/Cl ratios are generally lower than 1 both for springs and groundwater
 419 (Fig. 6a), with few exceptions such as the two thermal springs of 2018 sampling, and the
 420 CP53 and CP57 samples that show Na/Cl ratio higher than 1.

421 Spring waters mostly lie on the 1:1 line in Fig. 6b, indicating that calcium and magnesium
 422 originate from the dissolution of calcite, dolomite, and gypsum/anhydrite. Downstream
 423 groundwater, if compared to the 1:1 line, shows either excess (CP57, CP53, and some
 424 springs of 2018 sampling) or depletion of sulphate and bicarbonate. The former are Na-
 425 HCO_3 and NaMixCl, NaCl, and NaMix SO_4 waters respectively, while the latter (that also
 426 show Na/Cl ratio lower than 1) could be associated with the effects of dissolution of halite
 427 followed by inverse ion exchange, with calcium and/or magnesium increase and sodium
 428 depletion (Fig. 7a).

429 On the contrary, CP53 and CP57 HFs indicate a direct ion-exchange (Na increase and
 430 calcium and magnesium depletion) leading to Na enrichment. This means that the
 431 leaching of evaporite minerals such as halite, gypsum, carbonates, natron, and trona,
 432 typically observed in Algerian chotts (Zatout et al., 2020) is coupled with significant cation
 433 exchange processes in which Na is exchanged with Ca and vicariant elements.

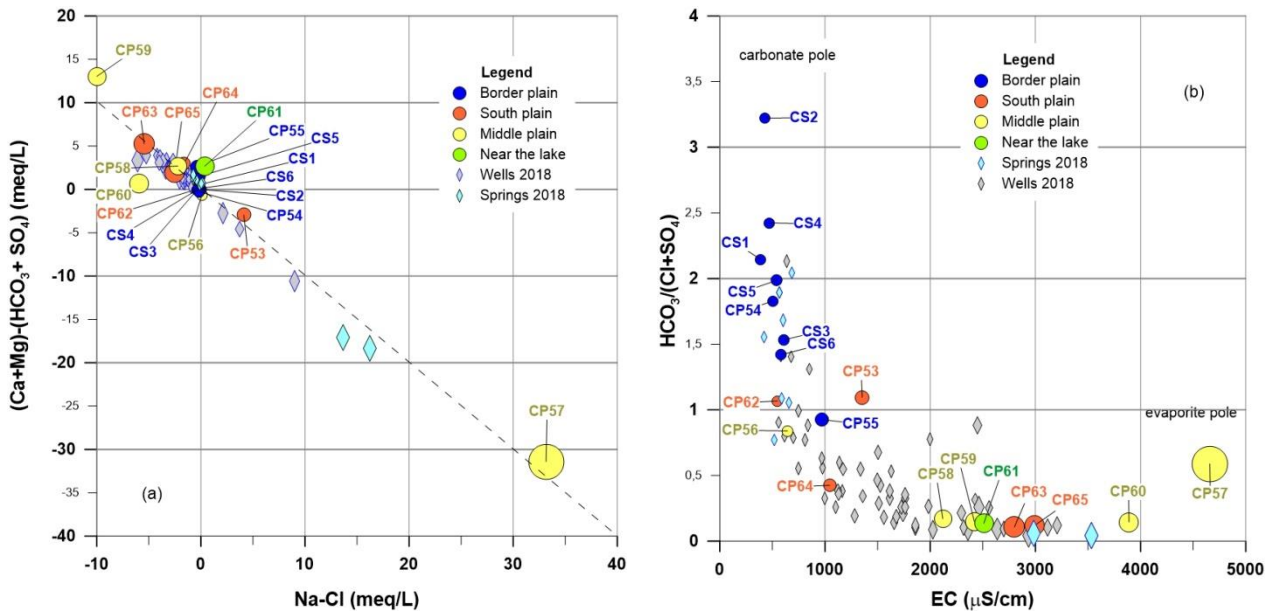
434 Fig. 7b shows the relationship between the $\text{HCO}_3/(\text{SO}_4+\text{Cl})$ ratio and EC; it confirms that
 435 the EC increase from the left (carbonate pole) towards the right (evaporite pole) is mostly
 436 due to the increase of both chlorides and sulphates. The samples related to CP53 (Na-
 437 HCO_3), CP57 (NaMixCl), two thermal springs (NaCl and NaMix SO_4), and CP59 (CaCl)
 438 deviate from the general trend because of the overlap of cation-exchange (direct or
 439 inverse).
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Fig. 6. (a) Relationship between Na and Cl concentrations; (b) Relationship between Ca+Mg and SO_4+HCO_3 concentrations. In both graphs, the symbol size is proportional to TDS.

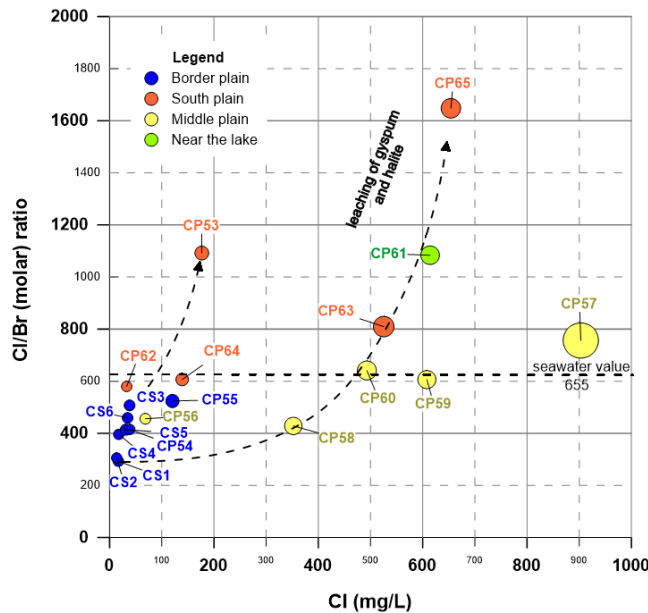
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Fig. 7. a) Relationship between the $(Ca+Mg)-(HCO_3+SO_4)$ difference and the Na/Cl ratio; (b) relationship between the $HCO_3/(SO_4+Cl)$ ratio and EC. The symbol size is proportional to TDS.

451 Fig. 8 shows the relationship between the Cl/Br (molar) ratio and chlorides. The Cl/Br
452 ratios of the low salinity springs should reflect the ratio of the local rainfall. Evaporation of
453 infiltrating water increases the chloride concentration in recharge waters, but leaves Cl/Br
454 unaffected: thus recharge by direct rainfall infiltration with subsequent evaporation seems
455 to interest most of the ground-waters of the Middle plain. The most likely reason for the
456 concurrent increase of Cl/Br ratios and salinity is the dissolution of halite that is contained
457 in the local salty lithologies.



458

Fig. 8. Cl/Br (molar) ratio vs. chloride concentrations; the symbol size is proportional to TDS.

459 Lithium (Fig. 3Sa, Supplementary Material) and Boron (Fig. 3Sb, Supplementary Material)
460 tend to increase, although not systematically, in parallel with the Cl/Br, suggesting that
461 boron minerals (borax, colemanite, ulexite; Helvacı and Palmer, 2017) and lithium minerals

462 (Li-rich phyllosilicates, Mertineit and Schramm, 2019; Bekele and Schmerold, 2020; Zatout
463 et al., 2020) are locally present in the salt pans and can be leached; then a simultaneous
464 cation exchange occurs (Abdelkader et al., 2012; Demdoum et al., 2015).

465 Among trace elements that can vicariate calcium, Sr (probably hosted in sulphates and
466 carbonates) is preferentially released compared to Ba in parallel with salinization, as
467 demonstrated by the positive correlation between the Cl/Br and Sr/Ba ratios (Fig. 4S).

468 Most of the Cl/Br ratios below seawater value are associated with medium and high nitrate
469 concentrations. Fig. 5Sa (Supplementary Material) indicates possible localized
470 groundwater pollution caused by the use of Br-based pesticides in agriculture and leaching
471 of septic waste (Alcalá and Custodio 2008). CP53 and CP61 show nitrate concentrations
472 lower than those found in the springs, close to background levels.

473 Fig. 5Sb (Supplementary material) shows the relationship between NO₃ and TDS. Two
474 main groups can be distinguished in the more saline waters: (A) with high nitrate contents
475 and (B) with low nitrate contents. As to Group A, nitrate concentrations indicate input from
476 irrigation return flows containing nitrogen compounds: CP63 and CP65 also show high
477 Cl/Br ratios. Group B shows nitrate concentrations less than 40 mg/L, more likely
478 influenced by septic tanks leaching. In this view, it is important to emphasize that recent
479 studies in salty environments demonstrated that nitrogen retention is microbially mediated
480 and can be stocked and released (Yang et al., 2015).

481 Fortunately, salinization and related cation exchange processes do not release in
482 groundwater toxic heavy metals such as Ni, Co, Cr, V, Cu, Zn, Pb, As, and U (Table 4S,
483 Supplementary Material) that often are below the detection limits and always below
484 common WHO thresholds defined for drinking water.

485

486 4.3 Isotopic composition of the studied waters

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488 Oxygen and hydrogen isotopes were also measured for groundwater and spring in the
489 F'Kirina plain (Table 2). The waters have $\delta^{18}\text{O}$ that varies between -9.7 ‰ and -5.9 ‰, with
490 an average value of -7.1‰, while δD varies between -61.3‰ and -40.3‰, with an average
491 value of -50.1‰. The springs located at the border of the plain show a $\delta^{18}\text{O}$ ranging
492 between -8.1‰ and -7.3‰ with an average value of -7.6‰, and a δD ranging between -
493 49.1‰ and -45.1‰, with an average value of -47.8‰. Isotopic data are represented in the
494 $\delta^{18}\text{O}$ - δD binary plot compared to the Global and Local Meteoric Lines (GMWL, Craig,
495 1961; Gat and Dansgaard, 1972).

496 The d-excess values were computed assessing potential evaporation effects ($d\text{-exc} = \delta\text{D} -$
497 $8 \cdot \delta^{18}\text{O}$, Dansgaard, 1964). Large variability is observed across the global and local scales.
498 The Eastern Mediterranean (Gat and Dansgaard, 1972; Bowen and Revenaugh, 2003)
499 shows values higher than 15, with a gradient of 6‰ from the Aegean Sea to the coast of
500 Israel (Saighi, 2001). The meteoric line for the Eastern Mediterranean is estimated to be
501 on average $\delta\text{D} = 8 \cdot \delta^{18}\text{O} + 20$ (Saighi, 2001; Bowen and Revenaugh, 2003; Aouad-Rizk et
502 al., 2005) while for the Western Mediterranean it is $\delta\text{D} = 8 \cdot \delta^{18}\text{O} + 13.7$ (Celle-Jeanton et al.,
503 2001). This reflects the difference of origin and the vapour supply and removal history of
504 the air masses over the two areas, which is also observed in recent isotopic data of
505 atmospheric vapour. In this study, the d-excess varies between 2.3‰ and 16.7‰ for
506 groundwater and between 9.3‰ and 16.7‰ in the spring waters.

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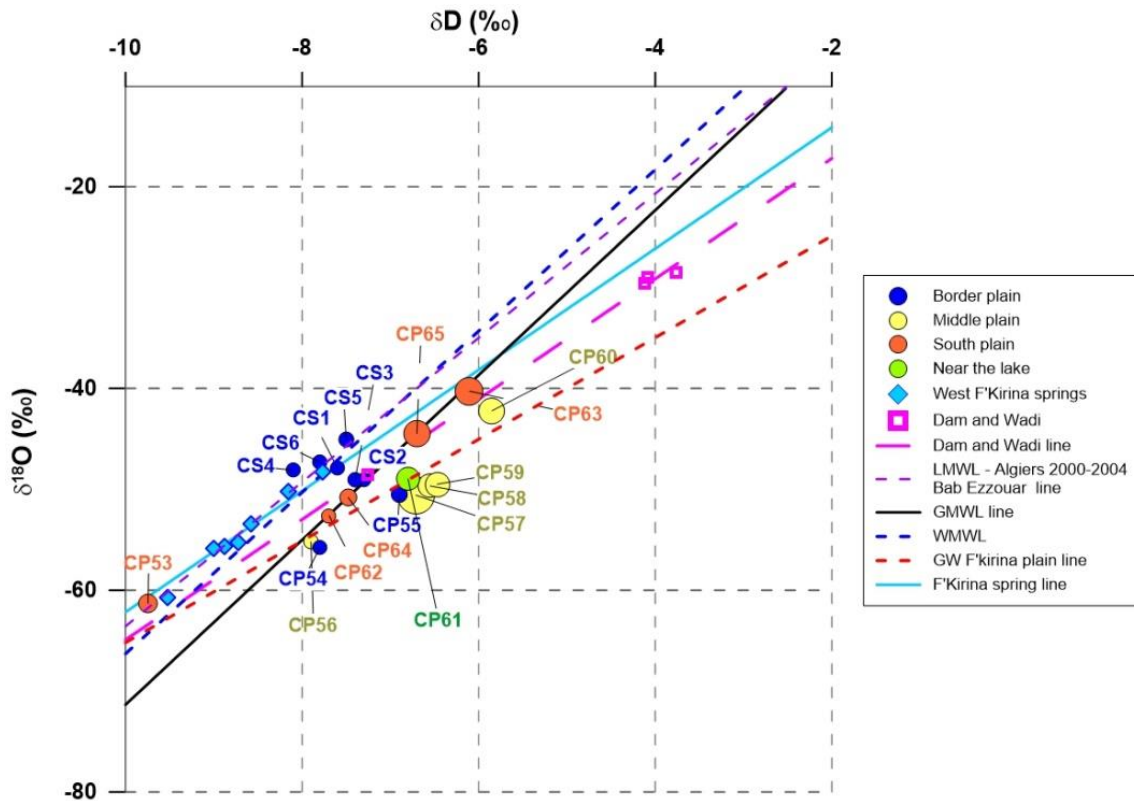
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Table 2. Stable isotopes of the F'Kirina plain waters.

Sample code	δD (‰)	$\delta^{18}O$ (‰)	d-exc (‰)
<i>Border plain</i>			
CS1	-47.9	-7.6	12.9
CS2	-49.1	-7.3	9.3
CS3	-49.0	-7.4	10.2
CS4	-48.1	-8.1	16.7
CS5	-45.1	-7.5	14.9
CS6	-47.3	-7.8	15.1
CP54	-55.8	-7.8	6.6
CP55	-50.5	-6.9	4.5
<i>Middle plain</i>			
CP56	-55.2	-7.9	8.1
CP57	-50.6	-6.7	8.1
CP58	-49.7	-6.6	3.1
CP59	-49.5	-6.5	2.7
CP60	-42.2	-5.9	2.3
<i>South plain</i>			
CP53	-61.3	-9.7	16.7
CP62	-52.6	-7.7	8.9
CP63	-40.3	-6.1	8.6
CP64	-50.8	-7.5	9.0
CP65	-44.5	-6.7	9.0
<i>Near the lake</i>			
CP61	-49.0	-6.8	4.6

514

515 Fig. 9 shows the $\delta^{18}O$ - δD relationship of the sampled waters together with *i*) the Local
516 Meteoric Water Line (LMWL) referred to isotope data of precipitation collected at USTHB
517 University in Bab-Ezzouar (Algiers) between 2000-2004 ($\delta D = 7.15 \delta^{18}O + 7.92$, $R^2 = 0.9$,
518 $n=113$; Saighi 2005), *ii*) the Western Mediterranean Water Line (WMWL, $\delta D =$
519 $8 * \delta^{18}O + 13.7$, Celle-Jeantone et al 2001), and *iii*) the Global Meteoric Water Line (GMWL,
520 $\delta D = 8.17 * \delta^{18}O + 10.35$, Rozanski et al 1993). Fig. 9 also shows the isotope stable
521 composition of springs of high altitude emerging from Aures mountains, located to the
522 south of F'Kirina plain (data from Belkoum et al 2020), and of waters from wadis and dams
523 of Chemora area, which locates to the west of F'Kirina plain (data from Belkoum and
524 Houha 2017). Springs of the Aures Mountains and springs of the Border plain group have
525 been merged for reconstructing the F'Kirina spring line. The GW F'Kirina Plain line
526 represents the Linear Regression Line of stable isotope data of the groundwater samples
527 from the wells of the F'Kirina plain.



528 Fig. 9. $\delta^{18}\text{O}$ - $\delta^2\text{D}$ diagram concerning the groundwater (GW) and spring waters of the F'Kirina area. Symbol
 529 size is proportional to the TDS of samples.

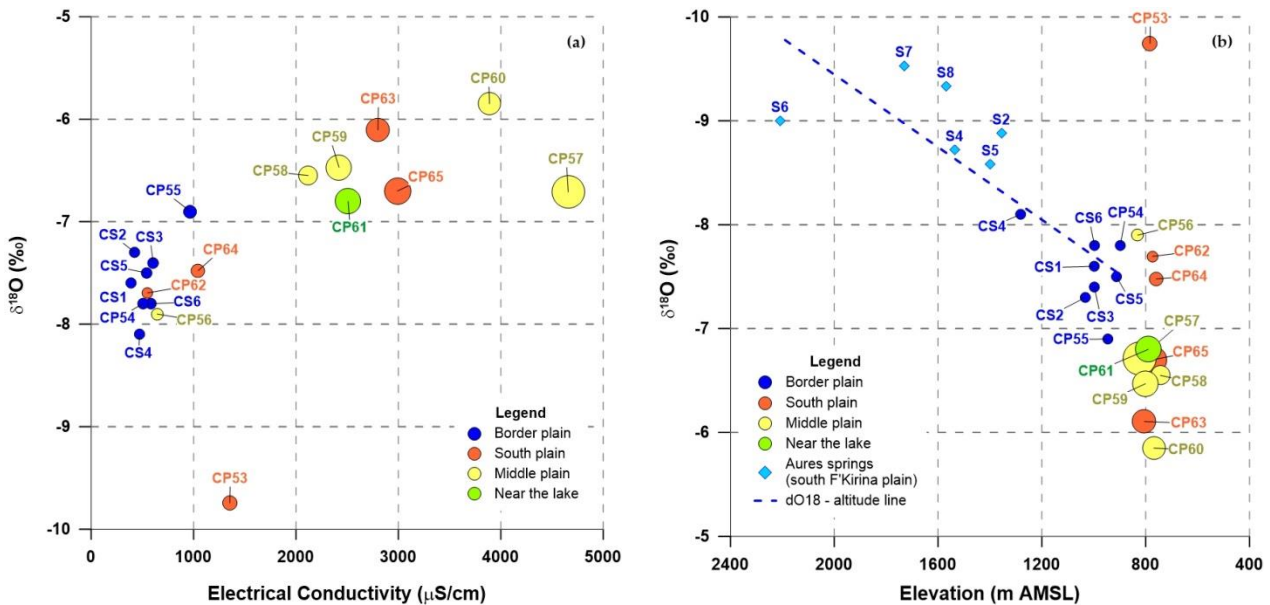
530 Isotope data of sampled waters show a slight positive correlation ($R^2=0.57$, n. 19) with a
 531 regression line having a slope of 4.2, which is quite different compared to the slope of 8
 532 related to the WMWL. The springs collected at the plain border are mostly located close to
 533 the WMWL and LMWL, indicating the possible contribution of local precipitation to the
 534 recharge. On the WMWL and LMWL, we also find the high elevation springs of the Aures
 535 massifs.

536 Most of the ground-waters of the plain are placed at the right of LMWL and GMWL. Their
 537 average d-excess, excluding the sample CP53 (d-excess 16.7‰), is 6.3‰: CP53 is the
 538 only groundwater sample placed close to the LMWL with more negative isotope values
 539 than those of both the springs of South and South-East mountains. If we exclude springs,
 540 the $\delta^2\text{D}$ - $\delta^{18}\text{O}$ relationship of the only ground-waters from wells (F'Kirina GW line in Fig. 9)
 541 shows a good correlation ($R^2=0.82$, n. 13), with an equation having a slope of 5. The line
 542 related to waters from dams and wadis of the Chemora area follows a trend mostly parallel
 543 to the F'Kirina GW line.

544 Groundwater samples from wells are aligned along a regression line; this can be explained
 545 by taking into consideration that during evaporation processes the isotopic composition of
 546 the residual liquid gradually departs from the LMWL, following a nearly linear trajectory
 547 known as evaporation line (Gat and Gonfiantini, 1981; Gonfiantini 1986).

548 If we consider the intersection of the F'Kirina GW and LMWL lines, the resulting isotopic
 549 value is close to the CP53 signature ($\delta^{18}\text{O}$ -9.4‰ and δD -61.7‰), providing useful
 550 information to estimate the isotopic composition of the recharging waters (Gibson et al.,
 551 1993). The isotopic composition of this water plausibly represents a recharge from the high
 552 parts of the Aures Mountains, and/or the contribution of past precipitation, with different
 553 climatic conditions compared to the present time (Tarki et al., 2012; Demdoum et al., 2015;
 554 Mokadem et al., 2016). In other words, we assume that CP53 and the spring samples (that

555 reflect the average isotopic composition of the current precipitation) are the end-members
 556 of the recharge process that variously mix to feed the groundwater system of the plain.
 557 This framework is coherent with the $\delta^{18}\text{O}$ vs. electrical conductivity and $\delta^{18}\text{O}$ vs. elevation
 558 relationships observed in Figs. 10a and 10b respectively. In particular, Fig. 10a shows that
 559 when EC exceeds 2000 $\mu\text{S}/\text{cm}$ the $\delta^{18}\text{O}$ values concurrently increase. Fig. 10b shows the
 560 relationship between $\delta^{18}\text{O}$ and elevation for springs emerging both from the Aures
 561 Mountains, located south of F'Kirina plain, and those belonging to the South-East
 562 mountain range. The regression of $\delta^{18}\text{O}$ vs. elevation data for the whole springs gives an
 563 altitude gradient of $-0.27 \delta^{18}\text{O}/100 \text{ m}$.
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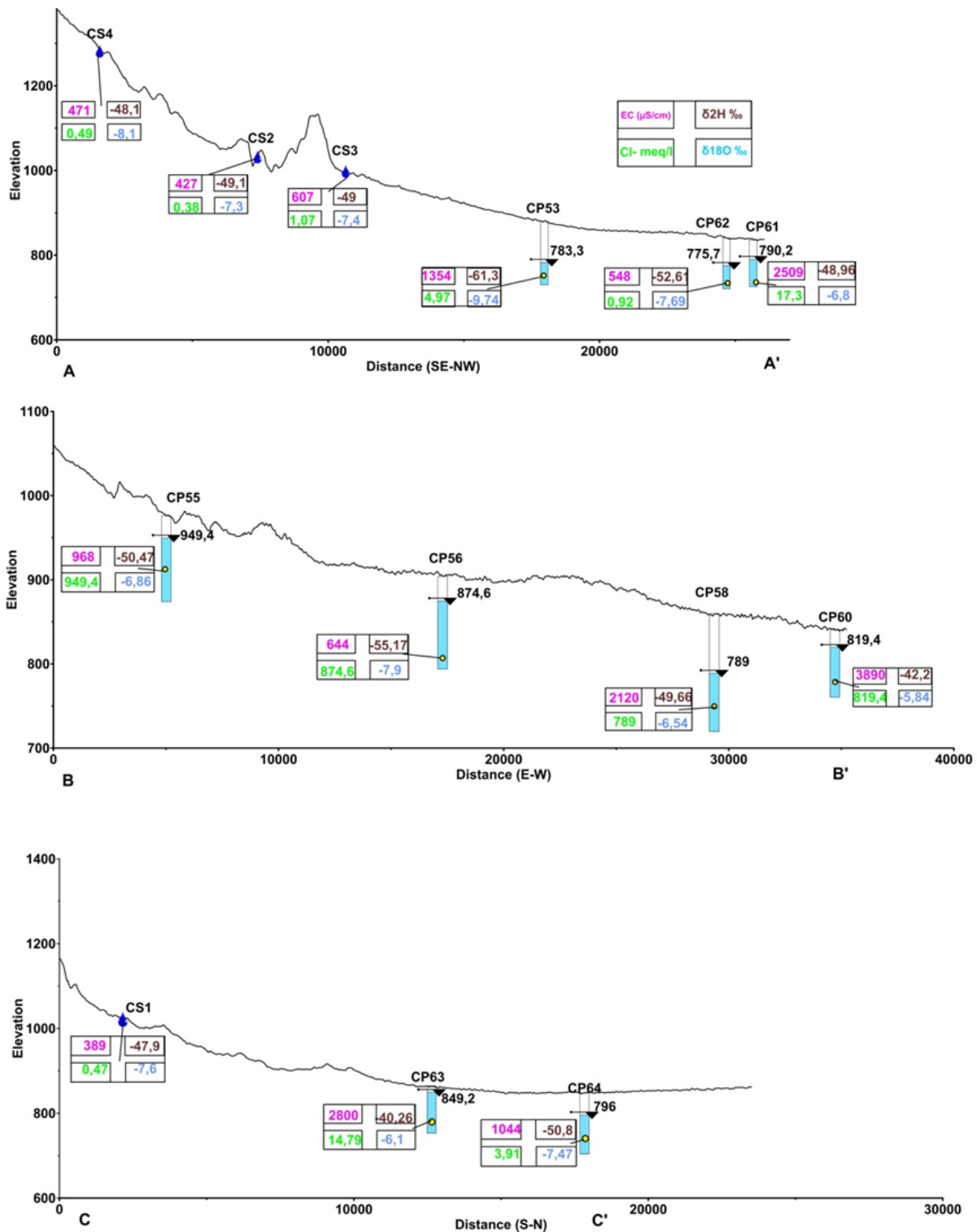
565 Fig. 10. (a) Relationship between $\delta^{18}\text{O}$ contents and EC; (b) Relationship between $\delta^{18}\text{O}$ and elevation of
 566 springs and well-heads. Symbol size is proportional to sample chloride concentrations.

567 It is also interesting to evaluate the stable isotope variation along some selected transects
 568 traced from the border of F'Kirina plain towards the lake (traces in Fig. 1). The transect AA'
 569 in Fig. 11 (with direction SSE-NNW) connects the springs CS4, CS2 and CS3, and the
 570 wells CP53, CP62 (South plain) and CP61 (Near the lake) (Table 2).

571 The $\delta^{18}\text{O}$ does not increase systematically towards the lake, while the water level in wells
 572 decreases; however, CP61 shows a higher water level than the upstream wells. CP53
 573 does not connect with stable isotope data of the considered springs: in the light of Fig.
 574 10b, its isotopic composition should refer to a recharge area at about 1700 m AMSL. CP62
 575 too shows a depleted isotopic composition compared to the springs located along the AA'
 576 transect, while only CP61 has an isotopic composition similar to that of springs.

577 The transect BB' (with direction E-W, Fig. 11) intercepts the samples: CP55, CP56, CP58,
 578 and CP60. The $\delta^{18}\text{O}$ becomes less negative towards the lake going from CP55 to CP58,
 579 and CP60; however, CP56 is out of the trend, having more depleted isotopic values and a
 580 saline content lower than the upstream groundwater (CP55). CP60 shows both a higher
 581 water level and higher salinization than the upstream well. The transect CC' (directed SE-
 582 NW, Fig. 11) crosses CS1 spring and CP63, CP64, and CP65 wells.

583 Even here we observe that CP64 has isotope values lower than CP65 located upstream.



584 Fig. 11. Water level, geochemical and isotopic variation along the AA', BB' and CC' transects (traces in Fig.
 585 1).
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587 Taken together, isotope data along transects suggest that water samples do not represent
 588 groundwater of a unique aquifer. We should remind that sampled wells (whose equipment
 589 and position of pump is unknown) can intercept different saturated levels, locally
 590 disconnected by less permeable material; however, wells can interconnect saturated
 591 levels, giving a pumping sample with “mixed information”, which is not easy to unravel.
 592 Besides this possible saturated level mixing due to well artefacts, we should also consider
 593 the contribution of return flow, which effects overlap those of the complexity of the aquifer

594 saturated levels.

595 **5. Conclusions**

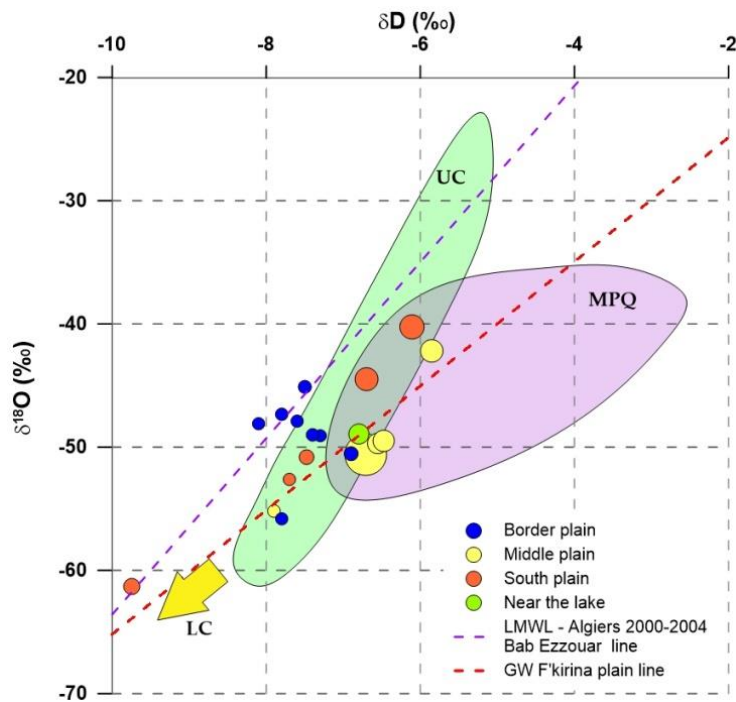
596 The hydrogeological system that feeds groundwater in the F'Kirina plain is complex as
597 made by interconnected aquifers represented by Mesozoic, Cenozoic, and Quaternary.
598 Piezometric investigation shows that a groundwater drought affected the F'Kirina plain
599 during the last 15 years likely because of the concurrence of a meteorological drought
600 succession and over-exploitation. Geochemical analyses indicate that the decreasing
601 water resource is also suffering from general salinization, mainly due to evaporation and
602 leaching of soil salts, a process that is coupled with simultaneous cation-exchange effects.
603 In this framework, we observe a general geochemical evolution from the fresh Ca-HCO₃
604 facies typical of springs bordering the plain towards more saline groundwater,
605 characterized by chloride/sulphate-rich facies, which are found in wells in the middle of the
606 plain, approaching the sebka. Uprising of deep saline groundwater, although subordinate,
607 is also recorded as suggested by the site CP57 and the neighbouring thermal springs.

608 Coherently, the oxygen and hydrogen isotopic analyses plotted in the notional $\delta^2\text{H}$ - $\delta^{18}\text{O}$
609 diagram record distinct trends, spanning between the regression line defined by the local
610 spring compositions, which reflects the characteristics of the water recharge components,
611 and the regression line of groundwater in the plain, which records various degrees of
612 evaporation.

613 Isotopic analyses also suggest that the F'Kirina plain ground waters incorporate far-field
614 water contributions that, having comparatively more negative oxygen and hydrogen
615 isotopic composition (δD - $\delta^{18}\text{O}$ down to -61.3‰ and -9.7‰, respectively), demonstrate
616 either their origin from altitudes higher than those of the sampled springs (δD between -
617 49.1‰ and -45.1‰, $\delta^{18}\text{O}$ between -8.1‰ and -7.3‰) or the drainage of old water
618 components recharged under climatic conditions cooler than at present.

619 These statements are corroborated by the isotopic diagram of Fig. 12, where the F'Kirina
620 plain groundwater composition is compared with that of groundwater of neighbouring
621 regions where the fingerprint of distinct aquifers has been highlighted. It can be noted that
622 F'Kirina plain groundwater mainly conforms to isotopic features of Upper Cretaceous
623 aquifers that are an important reservoir at the regional scale. The unique sample having
624 decidedly more depleted (negative) isotope ratios has been observed at the border of the
625 plain, where the older rocks, which are included in the syncline system, outcrop. This
626 composition recalls that typical of older aquifers that in surrounding areas are ascribed to
627 the Lower Cretaceous (Abid et al., 2011; Tarki et al., 2012; Kamel, 2013; Mokadem et al.,
628 2016). Deeper aquifers are generally characterized by more negative isotope composition
629 inherited by palaeo-precipitation having δD - $\delta^{18}\text{O}$ lighter than the present-day weighted
630 mean value for rain (Edmunds et al., 2003), possibly having Holocene (Abouelmagd et al.,
631 2012) or even Late Pleistocene age (Guendouz and Moulla, 2010; Darling et al., 2018).
632 However, the involvement of such deep and old water components cannot be assumed as
633 the general source of salinization, which appears mainly to be related to water-rock
634 interactions within the shallower (and salt-bearing) Miocene-Pliocene and Quaternary
635 aquifers in the presence of concomitant evaporation.

636 The comparison of current results with those of other surveys carried out in the F'Kirina
637 plain shows the existence of differences in groundwater chemical composition in distinct
638 years. These differences may indicate that the aquifer system, in its most exploited part,
639 has a flow connection of local or intermediate dimension, very sensitive to seasonal and
640 inter-annual variations, while some hydrochemical facies, which seem unusual compared
641 to the whole set, might be the expression of intermediate or regional flow systems (with a
642 deeper circulation).



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Fig. 12: Isotope composition of F'Kirina Plain waters compared to groundwater from Upper Cretaceous (UC), Lower Cretaceous (LC) and Miocene-Pliocene-Quaternary (MPQ) aquifers of surrounding localities. MPQ data from Tarki et al. (2012), Kamel et al. (2013) and Mokadem et al. (2016); UC data from Abid et al. (2011) and Mokadem et al. (2016); LC data from Abid et al. (2011) and Edmunds et al. (2003). GW is the regression line of F'Kirina area groundwater and the MWL is the local meteoric line of Algiers, as explained in section 4.5.

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Besides diffuse salinization, water levels and their comparison with previous information related to the last decade, suggest a decrease in groundwater volumes. Under possible climate changes that ultimately may lead to an increase in drought periods and a concomitant increase of water demand, groundwater shortage and salinization may probably get worse. The data collected in November 2019 represent the updated snapshot of the current conditions. They can represent an essential benchmark for future monitoring, to understand how the on-going climatic changes affect the local water resources in terms of both quantity and quality. The information is fundamental to provide advice and best practices for local water management.

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Author contribution

Conceptualization, R.O, M.D.F., G.L.; Methodology, M.C, N.C.; Investigation, R.O., G.L., M.C., N.C; Supervision, B.G., G.L., M.D.F., N.C.; Data collections, R.O.; Data curation, R.O, M.C., B.G., N.C., M.D.F.; Writing—Original Draft, R.O., M.C., M.D.F.; Writing—Review and Editing, B.G., N.C.. All authors have read and agreed to the published version of the manuscript.

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